

Assessment of Short-medium Term Intervention Effects Using CAE-SAR-Lisflood in Post-earthquake Mountainous Area

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Abstract. The 2008 Wenchuan earthquake triggered local geomorphic changes rapidly, producing abundant material through exogenic processes. The substantial material dynamics increased the risks of geo-hazards (flash floods, landslides, and debris flows) induced by extreme precipitation in the area. Intervention measures such as dams, levees, and vegetation revetments have been constructed in specified sites to reduce sediment transport, thus mitigating the risk of ensuing geo-hazards.

This study focused on assessing intervention effects incorporated with various facilities on post-earthquake fragile mountains in the short-medium term. Taking the Xingping valley as an example, we used the CAESAR-Lisflood, a two-dimensional landscape evolution model, to simulate three scenarios: unprotected landscapes, present protected landscapes, and enhanced protected landscapes between 2011 and 2013. We defined two indicators to assess the intervention effects of the three scenarios by comparing the geomorphic changes and sediment yield.

The results showed that the mitigation facilities were effective, especially engineering efforts cooperating with vegetation revetments in the upstream area. The spatial patterns of erosion and deposition changed considerably caused of the intervention measures. Additionally, the effectiveness of each intervention scenario showed a gradual decline caused directly by the reduction of the reservoir capacity. The enhanced scenario functioned better than the present one, with a smaller downward trend. The simulation results assessed the ability and effectiveness of cooperated control measures and could support optimum mitigation strategies.

1 Introduction

Strong earthquakes trigger co-seismic landslides and crack the mountains discontinuously, increasing weak structural planes (Huang, 2009) by weathering and erosion. Consequently, the source material produced from co-seismic landslides and attendant mass failure caused by the weak slope increase in mountainous regions and modify mountain landscapes by various surface processes for days, years, and millennia (Fan et al., 2020). The 2008 Wenchuan Ms 8.0 (the surface-wave magnitude, which is the logarithm of the maximum amplitude of ground motion for surface waves with a wave period of 20 seconds) earthquake has been influencing towns and other infrastructure in the affected area. Many studies have mapped the landslides triggered by the devastating earthquake. Gorum et al. (2011) performed an extensive landslide interpretation using a large set of high-resolution optical images and mapped nearly 60000 individual landslides, which are no less than 600 m². Xu et al. (2014) delineated 197481 landslides formed by polygons, centroids, and top points compiled from visual image interpretation. To estimate the threat of loose material in subsequent sediment disasters caused by landslides, some research attempt to measure the volume of deposited material based on field survey and assumptions. Huang and Fan (2013) estimated 400 million m³ of material deposited in the heavy-affected areas by assuming that the material deposited on steep slopes with angles larger than 30° and a catchment area of more than 0.1 km². An approximate 2793 million m³ of sediment was calculated by Chen et al. (2009) using different deposited depth settings in different buffer zones of the Longmenshan central fault. In summary, a

38 tremendous number of loose material accumulated on the gullies and hillslopes, ready to be eroded and transported away over
39 a long time. As a result, the mitigation in the Wenchuan quake-stricken area is still in the long run.

40 Structural mitigation measures have been developed in the affected areas regarding the site conditions and technical and eco-
41 nomic feasibilities. For example, ecological mitigation such as vegetation revetments was conducted to stabilise the source
42 area in hillslopes (Cui and Lin, 2013; Forbes and Broadhead, 2013; Stokes et al., 2014), and check dams were used widely to
43 intercept upriver sediment (Yang et al., 2021; Marchi et al., 2019). Lateral walls and levees were the longitude structures
44 (Marchi et al., 2019) to protect the infrastructures in mountain watersheds with higher sediment supply to the main streams.

45 Although comprehensive mitigation measures were performed in potentially dangerous sites, disasters still occurred owing to
46 rough terrain, vague source material, intensive precipitation, and relatively low-cost mitigating measures (Yu et al., 2010; Cui
47 et al., 2013). Therefore, understanding the effectiveness of intervention measures is crucial for mitigation strategies. Some
48 studies mainly focus on establishing post-evaluation effectiveness index systems that are not supported by sufficient practices
49 (Zhang and Liang, 2005; Wang et al., 2015). Some researchers compare the changes before and after the intervention measures
50 recorded by the long-term on-site measurement, which face the challenges to the investment of much time, energy and financ-
51 ing (Zhou et al., 2012; Chen et al., 2013). Recent research compares the disaster characteristics before and after the mitigation
52 actions, which are quickly obtained from the numerical simulation (Cong et al., 2019; He et al., 2022). Nevertheless, these
53 disaster characteristics express the process ignoring the long-time effects on the geomorphic changes (longer than the duration
54 of a single event). Therefore, the short-medium term (from the duration to decades) and spatial geomorphic changes quickly
55 obtained from the simulation will provide more details to interpret engineering measures in notable locations, even in those
56 inaccessible to humans.

57 CAESAR-Lisflood (C-L), which is based on the Cellular Automata (CA) framework (Coulthard et al., 2013), has powerful
58 spatial modelling and computing capabilities to simulate complex dynamic systems (Batty and Xie, 1997; Couclelis, 1997;
59 Coulthard et al., 2002). The model enables the study of many earth system interactions under different geo-environmental.
60 Representation of deposition and erosion within C-L is used widely in rehabilitation planning and soil erosion predictions from
61 a post-mining landform (Saynor et al., 2019; Hancock et al., 2017; J.B.C. Lowry et al., 2019; Thomson and Chandler, 2019;
62 Slingerland et al., 2019) and channel evolution and sedimentary budget with dam settings (Poepl et al., 2019; Gioia and
63 Schiattarella, 2020; Ramirez et al., 2020, 2022). In addition, there were a series of studies in the mountainous area involving
64 secondary geo-hazard driving factors (Li et al., 2018; Wang et al., 2014b) and vegetation recovery (Zhang et al., 2018). Li et
65 al. (2020) and Xie et al. (2018) have used C-L with different rainfall scenarios or future climate change to interpret the land-
66 scape evolution after the Wenchuan earthquake. The methods and parameter values used in the above research helped to pro-
67 mote the application in other study areas.

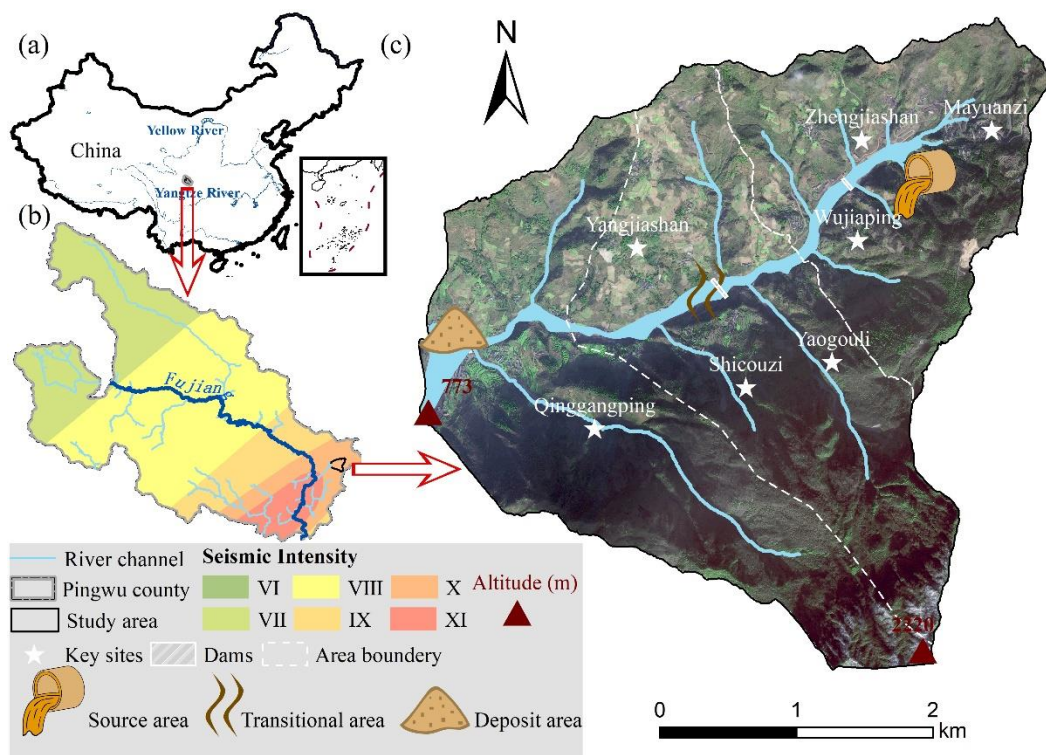
68 In this study, the hourly rainfall data of three years was generated by downscaling from daily to capture the extreme event.
69 Based on the input data, we simulated and compared the geomorphic changes and sediment yield in three scenarios that varied
70 in mitigation compositions and intensities in a catchment. The objectives are to 1) assess the effectiveness of a set of mitigation
71 facilities to reduce sediment transport, 2) analyse the role of each facility on geomorphic changes, and 3) determine vegetation
72 influence on catchment erosion.

73 **2 Study area**

74 **2.1 Regional characteristics**

75 The study area was Xingping valley in the northeastern Sichuan province, a left branch of the Shikan river (a tributary of the
76 Fu River) (Fig. 1). There are nearly two hundred households scattered among more than five villages in the catchment. The
77 topography of the catchment is rugged, with an elevation between 800 and 3036 m and an area of approximately 14 km². The

87 catchment is characterised by a high longitudinal gradient ($\sim 120\%$) and more than ten small V-shaped branch gullies. The
 88 length from northeast to southwest is 5770 m, the other direction perpendicular to which is 4150 m. A humid temperate climate
 89 with a mean annual temperature of 14.7°C characterises the region. The mean annual precipitation is 807.6 mm, mainly con-
 90 centrated between May and September. The steep terrain and short-term heavy rainfall make an ephemeral stream in this area.
 91 The local basement rocks are mainly metamorphic sandstones, sandy slate, crystalline limestone, and phyllite of Triassic
 92 Xikang Group ($T_{3\text{kk}}$) and Silurian Maoxian Group (S_{mx}), which are easily worn away by quickly weathering in a static process
 93 after disturbed in a strong earthquake. Consequently, the Wenchuan earthquake, with a Modified Mercalli Intensity scale of X,
 94 made this area one of the most severely affected locations (Wang et al., 2014a) and produced 10^6 m^3 loose material by triggering
 95 landslides and subsequent weathering in Mayuanzi, Zhengjiashan, and Wujiaping (Fig. 1)(Guo et al., 2018).



87
 88 **Figure 1: Overview of the study area. (a) Location of study area; (b) Seismic intensity map of the Wenchuan earthquake within the**
 89 **Pingwu county; (c) The schematic image of the study area.**

90 2.2 Historical hazards and intervention measures

91 Six group debris flow-flash flood disaster chains have been found in Xingping valley over a decade after the earthquake. Based
 92 on the published work of SKLGP (State Key Laboratory of Geohazard Prevention and Geoenvironment Protection) and the
 93 local states' geological survey before 2018 and our biannual field surveys since 2012, we catalogued the time of occurrence,
 94 total rainfall of each event, corresponding disaster details (Table S1). The massive sediment was transported quickly after the
 95 devastating quake in 2008 and 2009, and the extreme rainfall in 2013 and 2018 triggered prosperous loose material deposited
 96 in the channel. Considering the landslide processes, we divided the study area into three regions: source area, transitional area,
 97 and deposit area (the white dashed lines in Fig. 1c), which means the loose solid material would be easily transported from the
 98 source area to the deposit one through the transitional zone.

99 An engineering control project was constructed to intercept the upriver material in October 2010. The project included two
 100 check dams, one in the upper source area and the other in the transitional zone (Feng et al., 2017) (Fig. 1c). The upper dam
 101 has a storage capacity of $5.78 \times 10^4 \text{ m}^3$ and a height of 10.0 m. The transitional area dam has a storage capacity of $7.2 \times 10^4 \text{ m}^3$
 102 and a height of 9.0 m. With the reservoirs gradually filling with deposits, the first dredging work was subsequently done in

103 2013. Nearly three years later, the storage capacity behind the upper dam remained at 50% in 2016, while the transitional area
104 dam can no longer retain sediment.

105 3 Materials and Methods

106 In this study, we examined the intervention effectiveness through the morphological response and sediment yield in the Xing-
107 ping valley, which was simulated using the C-L model. The research entailed four main steps: 1) setting three scenarios with
108 different intervention compositions, 2) preprocessing the model input data, including three groups of DEMs, the rainfall data,
109 and m value of the C-L, 3) calibration of the hydrological component, and 4) simulating landscape changes and analysing the
110 intervention effectiveness in 2011-2013.

111 3.1 Scenarios settings

112 The abundant source material triggered by landslides should be controlled to prevent the threat of disasters downstream. There-
113 fore, we designed three scenarios by incorporating engineering and biological measures referenced to current facilities to assess
114 the effectiveness of intervention measures. Scenario UP: Unprotected landscapes meant the sediments would transport without
115 anthropogenic intervention. Scenario PP: Present protected landscapes implied that only the present two check dams trapped
116 deposits in 2011-2013 without dredging work over the period (see section 2.2). Scenario EP: Enhanced protected landscapes
117 emphasised the plus vegetation revetments in the source area and levees in the deposit area based on the two check dams in
118 Scenario PP.

119 Fig. 1c shows the locations of the existing two check dams in both Scenario PP and Scenario EP. We determined the placements
120 of additional facilities in Scenario EP according to the field survey, which demonstrated that the continuous supply of sediments
121 was mainly from the source area. Therefore, the vegetation revetments like tree planting would be carried out in upstream to
122 prevent erosion by stabilising topsoil and enhancing the soil's infiltration capacity with its roots (Lan et al., 2020).

123 Considering the damages caused by flash-flood to the residential area downstream, the levees (see Fig. S1 and Section 3.2.2)
124 are artificial barriers to protect agricultural land and buildings, which helped to prevent water and sediment from overflowing
125 and flooding surrounding areas. Table 1 shows the scenario descriptions, initial model conditions and input rainfall series. The
126 details about model input data are introduced below.

127 **Table 1: Scenarios setting**

Scenario	Descriptions	Period	DEM (10m)	Rainfall data
UP	no anthropogenic intervention		UP DEM UP bedDEM	downscaled hourly pre- cipitation in the period (lumped)
PP	the present two check dams upstream without dredging work additional vegetation revetments in the	2011-2013 (3 years)	PP DEM PP bedDEM	downscaled hourly pre- cipitation in the period (spilt)
EP	source area and levees in the deposit area based on Scenario PP		EP DEM EP bedDEM	

128

129 3.2 CAESAR-Lisflood

130 The C-L integrated the Lisflood-FP 2D hydrodynamic flow model (Bates et al., 2010) with the CAESAR landscape evolution
131 model (LEM) (Coulthard et al., 2002; Van De Wiel et al., 2007), which is described in detail by Coulthard et al. (2013). The
132 catchment mode of C-L was applied in this study, within which the surface digital elevation model (DEM), the bedrock DEM,
133 the grain size distribution, and a rainfall time series are required to simulate the sediment transport and geomorphic changes.
134 There are four primary modules within C-L operated as follows:

135 (1) a hydrological module generates surface runoff from rainfall input using an adaption of TOPMODEL (Topography based
136 hydrological model) (Beven and Kirkby, 1979),
137 (2) a hydrodynamic flow routing module based on the Lisflood-FP method (Bates et al., 2010) calculates the flow depths and
138 velocities,
139 (3) an erosion and deposition module uses hydrodynamic results to drive fluvial erosion by either the Einstein (1950) or the
140 Wilcock et al. (2003) equation applied to each sediment fraction over nine different grain sizes,
141 (4) and a slope model moves material from the hillslope to the fluvial system by considering both the mass movement when a
142 critical slope threshold is exceeded, and soil creep processes whereby sediment flux is linearly proportional to surface slope.
143 The C-L model updates variable values stored in square grid cells at intervals, such as DEM, grain size and proportion data,
144 water depth, and velocity. For three scenarios, the initial conditions, such as DEMs and bedrock DEMs, the rainfall data, and
145 the m values were preprocessed as follows.

146 **3.2.1 Surface and bedrock digital elevation model**

147 To describe clearly the control process, especially the two dams and levees in the catchment, we unified grid cell scales to 10
148 m for all input data of the C-L. The GlobalDEM product with a 10 m × 10 m resolution and 5 m (absolute) vertical accuracy
149 was used to form three types of initial DEMs (UP DEM, PP DEM, and EP DEM). Before rebuilding initial DEMs, we filled
150 the sinks of the original GlobalDEM based on Environmental Systems Research Institute's (ESRI's) ArcMap (ArcGIS, 10.8)
151 to eliminate the 'walls' and the 'depressions' in the cells and avoided the intense erosion or deposition in the early run time.
152 Then the non-sinks DEM was used as the surface DEM in Scenario UP (UP DEM) without any facilities. According to the
153 engineering control project described in Section 3.2.2, surface DEM of Scenario PP (PP DEM) included the dams by raising
154 the grid cell elevations by 10 m for the dam in upper stream and 9 m for the dam in the transitional area. Similarly, the surface
155 DEM in Scenario EP (EP DEM) included the dams in PP DEM. In addition, two levees were produced by raising grid cells'
156 elevation by 2 m that were represented at selected locations. For scenario EP, the placement and setting of vegetation revet-
157 ments in Scenario EP were introduced in Section 3.2.2.

158 The spatial heterogeneity of source material (Fig. 1c) indicates the discrepancy of erodible thickness, which equals the differ-
159 ence between surface DEM (DEM) and bedrock DEM (bedDEM). We divided the study area into five regions according to
160 the erodible thickness (Fig. S1) by checking out the relative elevation of the foundations of buildings, the exposed bedrock,
161 and the deposited depth of landslides to the ground level. The average thicknesses of upstream low and high-altitude areas
162 were set to 10 m and 3 m, respectively, and the thickness of erodible layer in the downstream area was set to 3 m. For the river
163 channel and outlet, there would be a large amount of deposition; the thickness of erodible sediment was set to 5 m and 4 m,
164 respectively. As the dams in Scenario PP and the levees in Scenario EP were non-erosive concrete, we set the erodible thickness
165 of these features to 0 m. Eventually, DEMs were formatted to ASCII raster as required by C-L. The divided regions varied in
166 erodible thickness, the placement of additional levees and vegetable revetments in Scenario EP, and the generation process of
167 DEMs and bedDEMs were shown in Fig. S1.

168 **3.2.2 Vegetation settings**

169 Another parameter required in each scenario simulation was the m value of hydrological model (TOPMODEL) within C-L,
170 which controls the exponential decline of transmissivity with depth (Beven, 1995, 1997) and influences the peak and duration
171 of the hydrograph in response to rainfall. The m value effectively imitates the effect of vegetation on the movement and storage
172 of water within the soil. The lower the m value, the lower the vegetation coverage, and the flashier flood peak and shorter
173 duration are reflected in the flood hydrograph (Coulthard et al., 2002). The m value is usually determined by the landcover
174 (e.g., 0.02 for the forest, 0.005 for the grassland) (Coulthard and Wiel, Van De J., 2017). In our study, we set the m value as

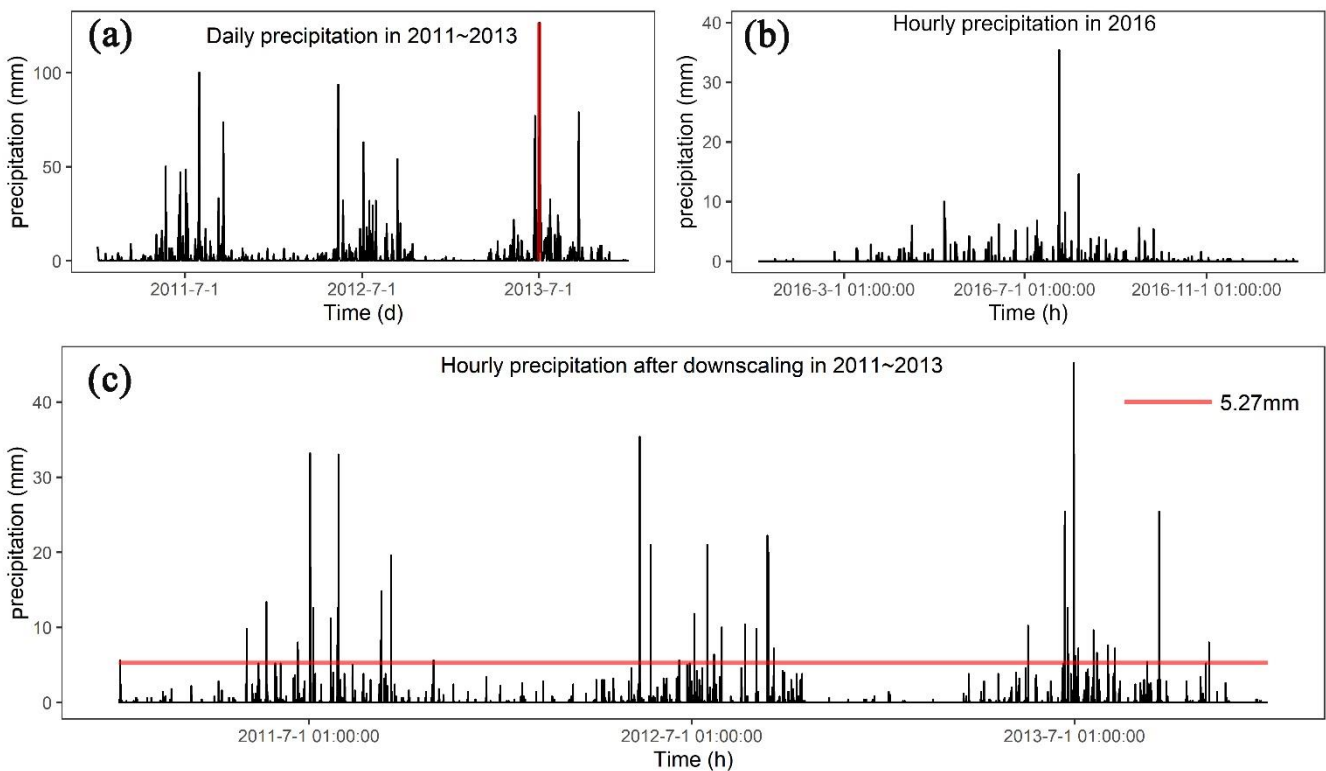
175 0.008 in our smaller catchment (14 km²) in Scenario UP and PP, which resembles the m value of farmland covered with lower
176 vegetation coverage in the same catchment studied by Xie et al. (2018) and Li et al. (2018). As mentioned earlier, the upstream-
177 low elevation area covered by the biological measures in the EP scenario was assigned a higher m value of 0.02. It has been
178 calibrated in the more extensive catchment containing our study area by replicating the flood event in 2013 (Xie et al., 2018).

179 3.2.3 Rainfall data

180 In this research, we compared three scenarios with identical precipitation data between 2011 and 2013, as mentioned in section
181 3.1. The source data of precipitation in 2011-2013 (Fig. 2a) was obtained from the China Meteorological Administration
182 (<http://data.cma.cn>) with daily temporal resolution. The intensity and frequency of extreme rainfall events affect patterns of
183 erosion and deposition (Coulthard et al., 2012b; Coulthard and Skinner, 2016). Therefore, we used the stochastic downscaling
184 method to generate hourly data to better capture the hydrological events introduced by Li et al. (2020) and Lee and Jeong
185 (2014). The referenced hourly precipitation was observed from the pluviometer located 20 km from the study area in 2016
186 (Fig. 2b), with an annual total precipitation of 684 mm. The observed rainfall in 2016 was characterised by: (1) hourly precip-
187 itation was from 1.1 mm to 35.4 mm, and (2) the maximum and average duration of a rainfall event was 24 h and 2.8 h,
188 respectively. The main processes of the downscaling method are:

- 189 ● extracting the hourly rainfall of specific days in 2016 closest to the daily rainfall in 2011-2013 through the threshold
190 setting and producing the genetic operators using the extracted hourly rainfall dataset;
- 191 ● mixing on the genetic operators by genetic algorithm (Goldberg, 1989) composed of reproduction, crossover and mutation
192 and repeating until the distance between the sum of hourly rainfall and the actual daily rainfall is less than the set threshold;
- 193 ● normalising the hourly precipitation to remain the daily rainfall value unchanged.

194 Fig. 2c shows the downscaled rainfall series between 2011 and 2013. The downscaled hourly rainfall better captured the hy-
195 drological events at an hourly scale compared to the hourly-mean rain (5.27 mm) on the day with extreme rainfall (126.5 mm),
196 which was far from the actual situation. Corresponding to the m value settings, the input of generated hourly precipitation is
197 catchment lumped in Scenario UP and Scenario PP and divided into two separate but identical rainfall in Scenario EP.



198

199 **Figure 2: (a) Daily precipitation in 2011-2013 (the red vertical line indicates maximum daily precipitation of 126.5 mm); (b) Hourly**
200 **precipitation in 2016; (c) Downscaled hourly precipitation in 2011-2013 (the red horizontal line indicates the hourly-mean precipi-**
201 **tation 5.27 mm on the day with maximum precipitation marked in (a)).**

202 **3.2.4 Other parameters**

203 The C-L is sensitive to a set of input data introduced by Skinner et al. (2018) for a catchment with a grid cell size of 10 m,
204 such as sediment transport formula, slope failure threshold, and grain size set. The grainsize distribution of sediment is derived
205 from samplings at 14 representative locations in the same study basin by Xie et al. (2018). Given the grainsize distribution in
206 this study, the Wilcock and Crowe formula was selected as the sediment transport rule, which was developed from flume
207 experiments using five different sand-gravel mixtures with grain sizes ranging between 0.5 and 64 mm (Wilcock et al., 2003).
208 Considering the steep slope on both sides of deep gullies, a higher slope failure threshold was determined to replicate the
209 geomorphic changes between 2011 and 2013. Additionally, we found that the probability of shallow landslides indeed accu-
210 mulated from 20° to 50° in slope gradient between 2011 and 2013 (Li et al., 2018). The slope angle was derivate from the DEM
211 with a 30 m spatial resolution, which caused a lower slope angle than that with a 10 m resolution. As such, we set 60°, which
212 is lower than the 65° used in a scenario without landslides (Xie et al., 2022) and higher than 50°. Some parameters were
213 determined by repeated experiments such as the minimum Q value, and the other input values were referred to default values
214 recommended by the developers (such as the max erode limit in the erosion/deposition module and the vegetation critical shear
215 stress) in <https://sourceforge.net/p/caesar-lisflood/wiki/Home/>. Table S2 in the supplemental material presented model param-
216 eters of C-L used in this study.

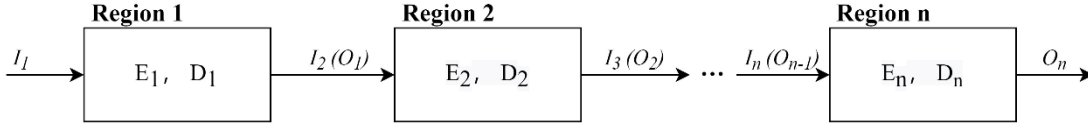
217 **3.2.5 Model calibration**

218 Considering the ungauged basins before 2015, we replicated the flash flood event in July 2018 using C-L simulations to cali-
219 brate the hydrological components. Based on Scenario PP (with two checking dams), we input the two-week hourly precipita-
220 tion in July 2018 (Fig. S2a), which is recorded by the rain gauge located 2.5 km away from the catchment (Fig. S2b). The
221 simulation results (Fig. S2c and Fig. S2d) show the erosion map and maximum water depth map in Scenario PP on July 15,
222 2018. We selected three locations to compare the deposition and inundation in simulation results, satellite images and photos
223 (Fig. S3). Additionally, the simulated sediment thickness and water depth were close to those measured from pictures, which
224 indicated that the flash flood event was well replicated by the C-L using the input data.

225 **3.3 Output analysis**

226 The C-L model outputs in each scenario include hourly water and sediment discharge at the basin outlet, the difference between
227 DEMs at a specified time and initial DEMs (EleDiffs). We validated the model outputs by comparing the hourly discharge and
228 EleDiffs reflecting the depth of sediment deposition or erosion (> 0.1 m: deposition, < -0.1 m: erosion) with field survey
229 materials. The overall temporal and spatial geomorphic changes reflected by EleDiffs under three different scenarios were
230 used to assess the geomorphic response to interventions. To explore the geomorphic response to various control measures, we
231 focused on the key spots placed checking dams, levees, and vegetation revetments and recorded the depth of deposited sedi-
232 ment behind two dams. To further explore the spatial heterogeneity, we compared the volumes of deposition and erosion among
233 three divided regions, including the source area, transitional area, and deposition area.

234 Based on the visual analysis and quantitative results, we defined two formulae to assess the effectiveness of intervention. The
235 conservation ability (Ca , Eq. (3)) was calculated based on variables in the sediment balance system (Fig. 3). The sediment
236 volume of deposited sediment (D_n) and input sediment from the upper connected region (I_n) is equal to that of eroded material
237 (E_n) and output sediment to the next part (O_n) over the same period (Eq. (1), Eq. (2)) in the system. A higher value of Ca in a
238 specific region and scenario indicates that a more effective control system is applied.



239
240 **Figure 3: The sediment balance system in the study area (the Region n indicated source area, transitional area, and deposit area)**

241

$$I_n = \sum_2^n E_{n-1} - \sum_2^n D_{n-1}, \quad (1)$$

$$I_n + E_n = O_n + D_n, \quad (2)$$

$$Ca = \frac{D_n}{I_n + E_n} \quad (3)$$

242 Where, n is the region number of source area (=1), transitional area (=2), and deposit area (=3).

243 Additionally, we designed the relative efficiency (Re , Eq. (4)) to depict the efficiency of intervention measures in Scenario PP
244 and EP in sediment loss, with the comparison to Scenario UP.

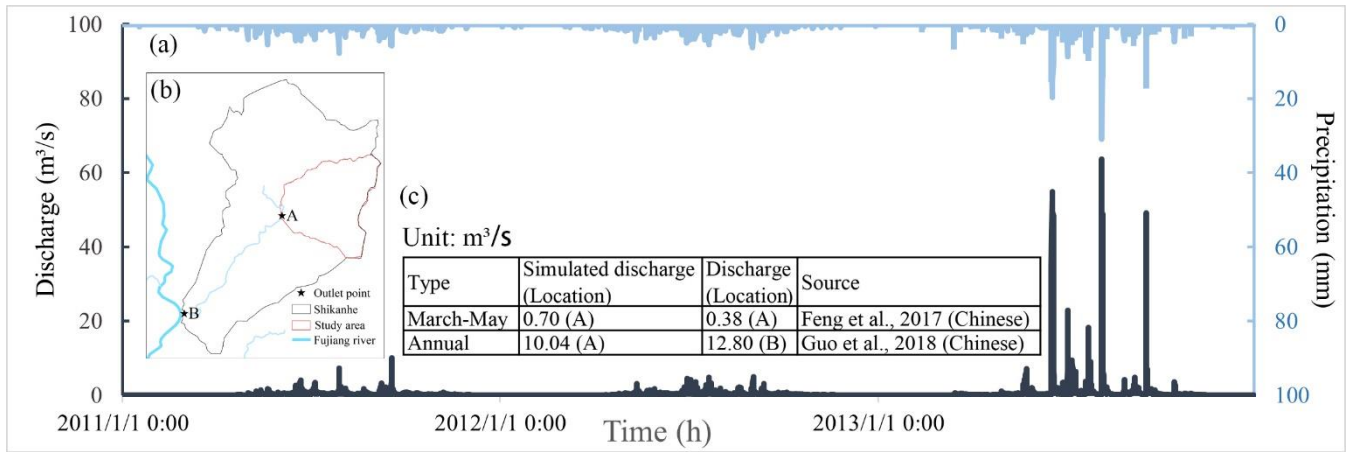
$$Re_{PP/EP,i} = \frac{Q_{UP,i} - Q_{PP/EP,i}}{Q_{UP,i}} \quad (4)$$

245 Where i is the sequence of the day; Q_{UP} is the daily sediment yield measured at the catchment outlet in Scenario UP; $Q_{PP/EP}$ is
246 the daily sediment yield measured at the catchment outlet in Scenario PP or Scenario EP of day i ; $Re_{PP/EP}$ is the daily relative
247 effectiveness of control measures in Scenario PP or Scenario EP.

248 4. Results

249 4.1 Model verification

250 Fig. 4 shows the input rainfall data and modelled discharge hydrograph between 2011 and 2013 (Fig. 4a). The comparison of
251 simulated mean discharge in April through July and the whole year with field survey materials in two locations are also pre-
252 sented (Fig. 4b, c). Concerning the discharge hydrograph, the peak discharges (63.7, 54.9, and 50.3 m³/s) correspond well with
253 the peak rainfall intensities (31, 19.7 and 15 mm). The modelled water discharge from March to May in location A is slightly
254 larger than the measured value reported by Feng et al. (2017). Additionally, an average annual discharge of 10.04 m³/s in
255 location A is less than that of 12.80 m³/s in the catchment outlet (location B), which has an area approximately three times the
256 size of the study area.



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Figure 4: The input and output of the hydrograph. (a) The input hourly precipitation and simulated discharge in 2011-2013 in Scenario PP; (b) Location of the specified outlet point; (c) the comparison of the simulated average discharge to the recorded discharge.

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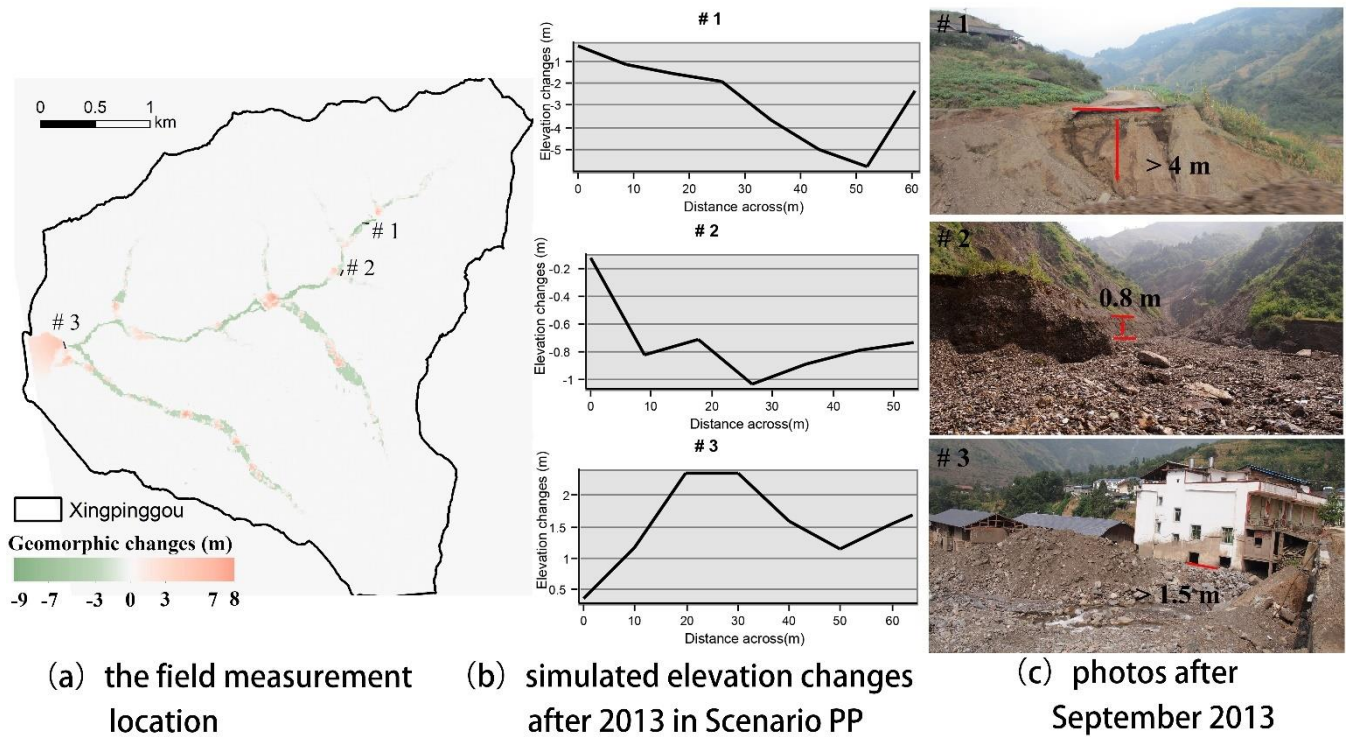
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Typical cross-sections are generated (Fig. 5) based on the replicated landscape changes in Scenario PP. The first site is located on the upriver road, which was eroded at a depth of 5.7 m according to the simulation results, while the photo shows a depth of no less than 4.0 m without an apparent eroded base. The cross-section #2 and site photo of the gully depict that the eroded depth is approximately 1.0 m. Meanwhile, a clear sediment boundary is found in the building located at the deposited area (# 3), indicating a slightly lower deposition depth than the modelled one.



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Figure 5: The comparison of cross-sections from the simulation results to the field measurements after 2013 in Scenario PP.

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4.2 Overall geomorphic changes

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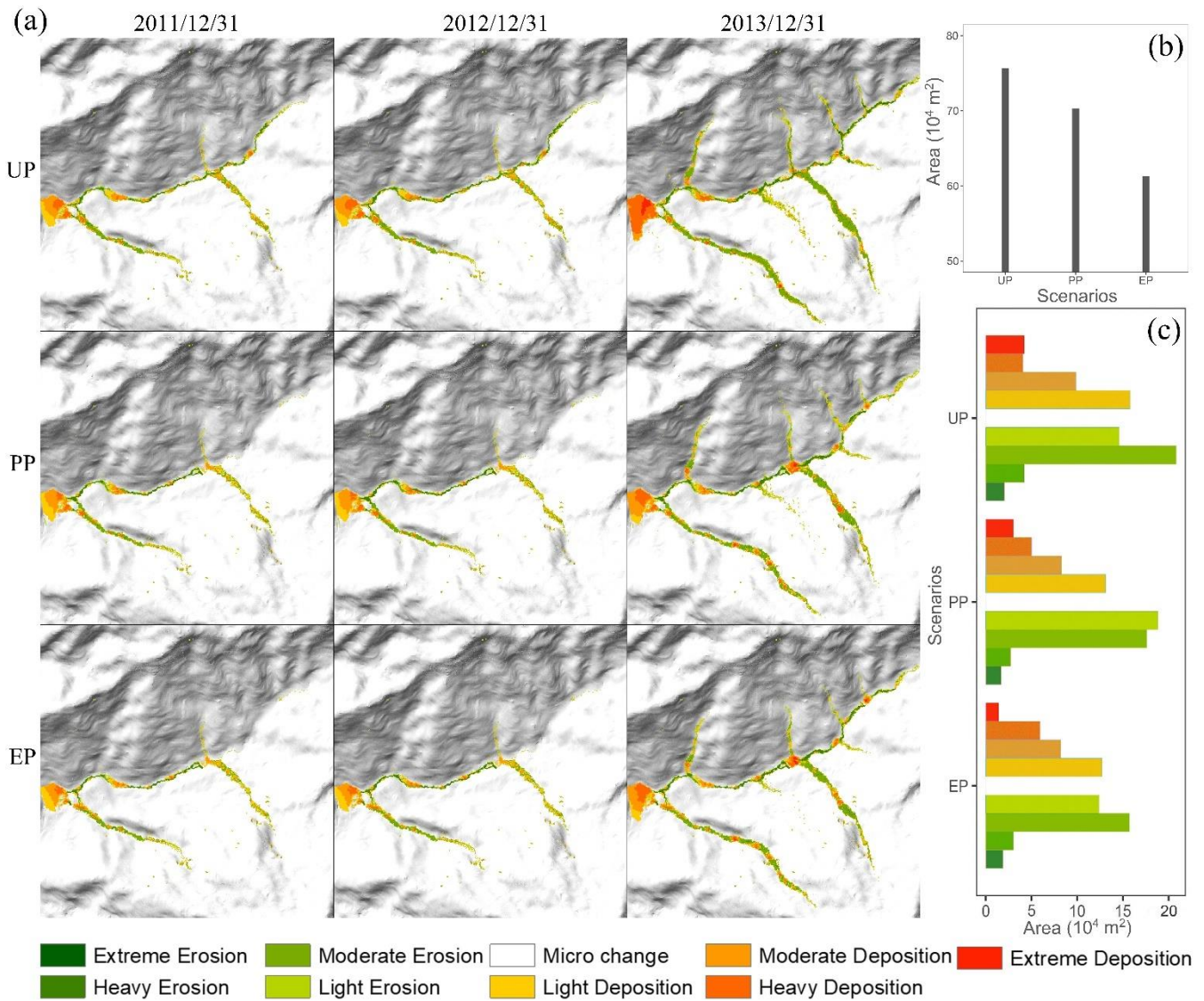
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Fig. 6a compares the three annual landscapes changes in each scenario, which were classified into nine categories by natural breaks for EleDiffs: extreme erosion (<-7 m), heavy erosion (-7--3 m), moderate erosion (-3--1 m), light erosion (1-0.1 m), micro change (-0.1-0.1 m), light deposition (0.1-1 m), moderate deposition (1-3 m), heavy deposition (3-7 m), and extreme deposition (>7 m). A similar spatial pattern of erosion is observed in all three scenarios. More specifically, erosion mainly

273 occurred in the main channel and the branch valleys, among which the left branches were more severe. In contrast, the depo-
 274 sition distribution appeared to be varied in three scenarios, especially the area behind the two dams shown in Scenario PP and
 275 EP.

276 The total area of affected grid cells representing erosion and deposition in each scenario was calculated to reveal the difference
 277 (Fig. 6b). The affected area in Scenario UP was the largest at about 0.76 km² (5.4% of the total catchment), which was larger
 278 than that in Scenario PP (0.70 km², 5.0% of the whole catchment), and the affected area decreased to 0.61 km² (4.4% of the
 279 total catchment) in Scenario EP. The total area of erosion and deposition reduced gradually with more controlling measures
 280 established in this study.

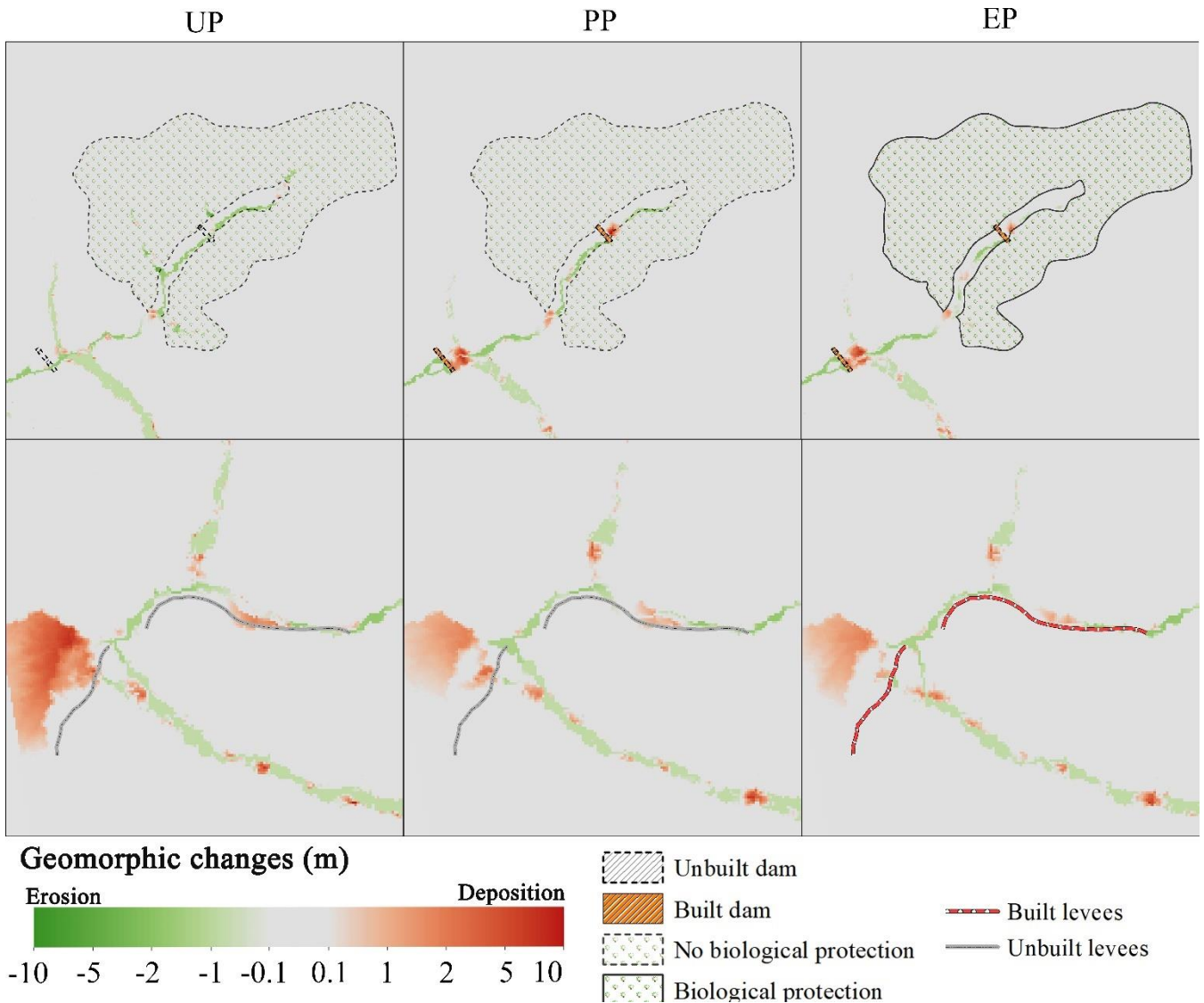
281 Fig. 6c compares the extent of geomorphic changes in three situations using the areas varied in depth. The light and moderate
 282 erosion areas were more than the extreme and heavy ones for all three scenarios. The size of each erosion degree in UP
 283 was more extensive than in PP and followed by EP. In addition, the greater the deposition depth, the less deposition coverage.
 284 Especially the extreme deposition area was somewhat more than the area of the heavy deposition in UP. Further analysis shows
 285 that extreme, moderate, and light deposition areas decreased in the order of UP, PP, and EP. The heavy deposition areas show
 286 the opposite trend, mainly attributed to the checking dams and vegetation revetments.



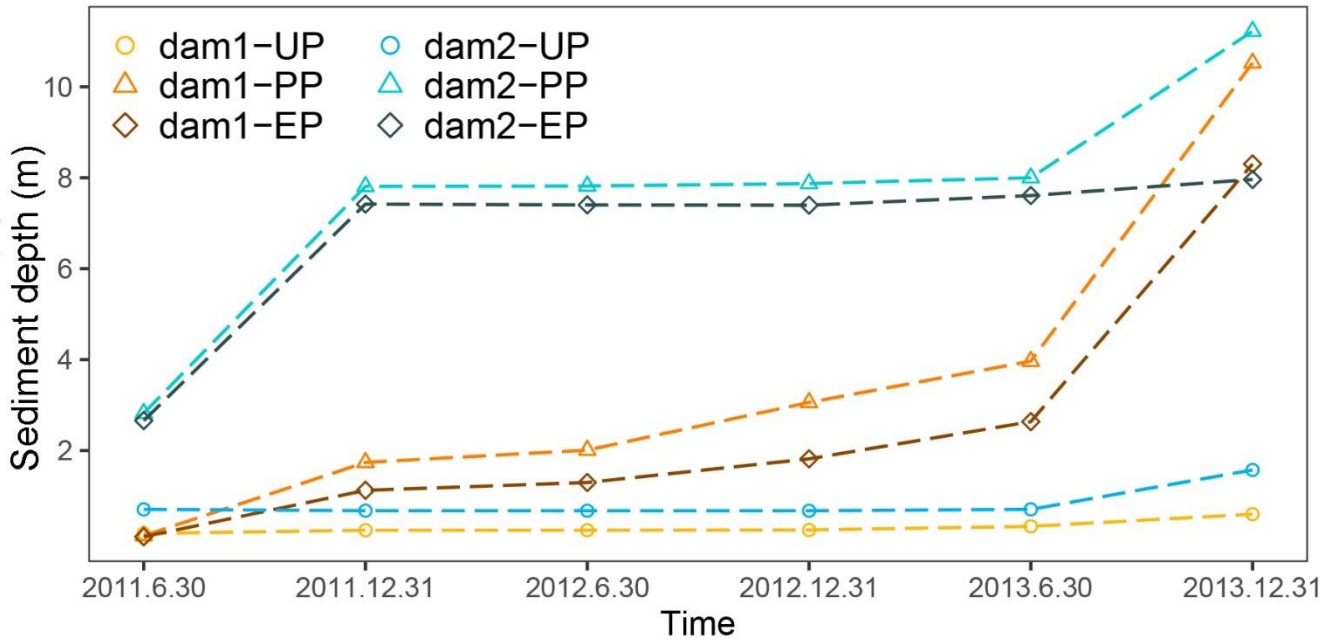
287
 288 **Figure 6: (a) Simulated geomorphic changes over time for three scenarios; (b) the affected area of deposition and erosion for three**
 289 **scenarios; (c) columnar distribution of different erosion and deposition levels.**

290 **4.3 Details of key spots**

291 As shown in Fig. 7, the controlling measures and surroundings for the three scenarios were further investigated. Behind the
 292 two dams upriver, the evident orange clusters indicate that the deposition occurred in Scenario PP and EP. In contrast, these
 293 locations were dominated by erosion, shown in green in scenario UP. Further analysis of the sediment depth shown in Fig. 8
 294 showed that the deposited depth behind the dams in Scenario EP was lower than those in Scenario PP. Additionally, in Scenario
 295 PP, sediment trapped by dam 1 was less than that by dam 2, but both have deposit thicknesses of more than 10 m, which
 296 exceeded the dams' heights (dam 1's height is 10 m, dam 2's height is 9 m). As for the simulation results in Scenario EP, the
 297 values of deposition depth behind the two dams were nearly 8 m, which were lower than the dams' heights.
 298 The material produced from upriver tributary gullies varied due to the additional biological protection measures in three sce-
 299 narios. A volume of $14.4 \times 10^4 \text{ m}^3$ sediments was transported from EP's biological protection area (solid lines in Fig. 7).
 300 $27.1 \times 10^4 \text{ m}^3$ and $16.9 \times 10^4 \text{ m}^3$ loose material were produced in the same region without biological protection in Scenario UP
 301 and PP, respectively. The vegetation revetment enhanced the sediment conservation based on the role of dam 1. Compared
 302 with the deposition in UP and PP without levees in the downriver area (shown in the bottom row of Fig. 7), the levees in EP
 303 blocked debris in the bend of the channel and played an essential role in protecting the residents and cultivated land behind
 304 the levees.



305
 306 **Figure 7: Geomorphic changes at key spots of the simulation results for the UP, PP, and EP scenarios. The top row is the upriver**
 307 **extent containing dam 1, dam 2 and the vegetation revetment. The bottom row is the downriver extent containing levees.**

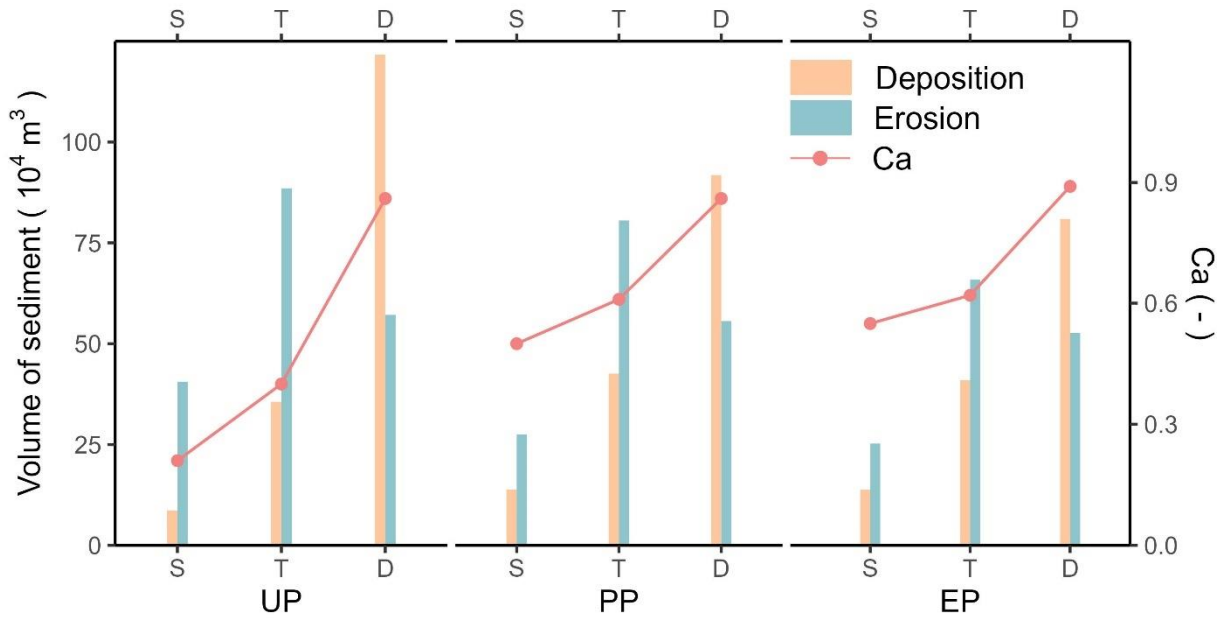


308
309 **Figure 8: The depth of deposited sediment in the dams' placements.**

310 **4.4 Effectiveness assessment of intervention measures**

311 Fig. 9 shows the erosion and deposition volumes in the source, transitional, and deposit areas and compares conservation
312 ability (Ca) in each scenario. For all three scenarios, the deposition volume in the source area was less than that in the transi-
313 tional area, and the largest amount of sediment was deposited in the deposit area. Regarding the eroded sediment, the largest
314 volume was in the transitional area, followed by the transitional area, and the source area presented the least volume. Moreover,
315 sediment transport was best controlled in the deposit area and worst contained in the source area under any intervention con-
316 ditions.

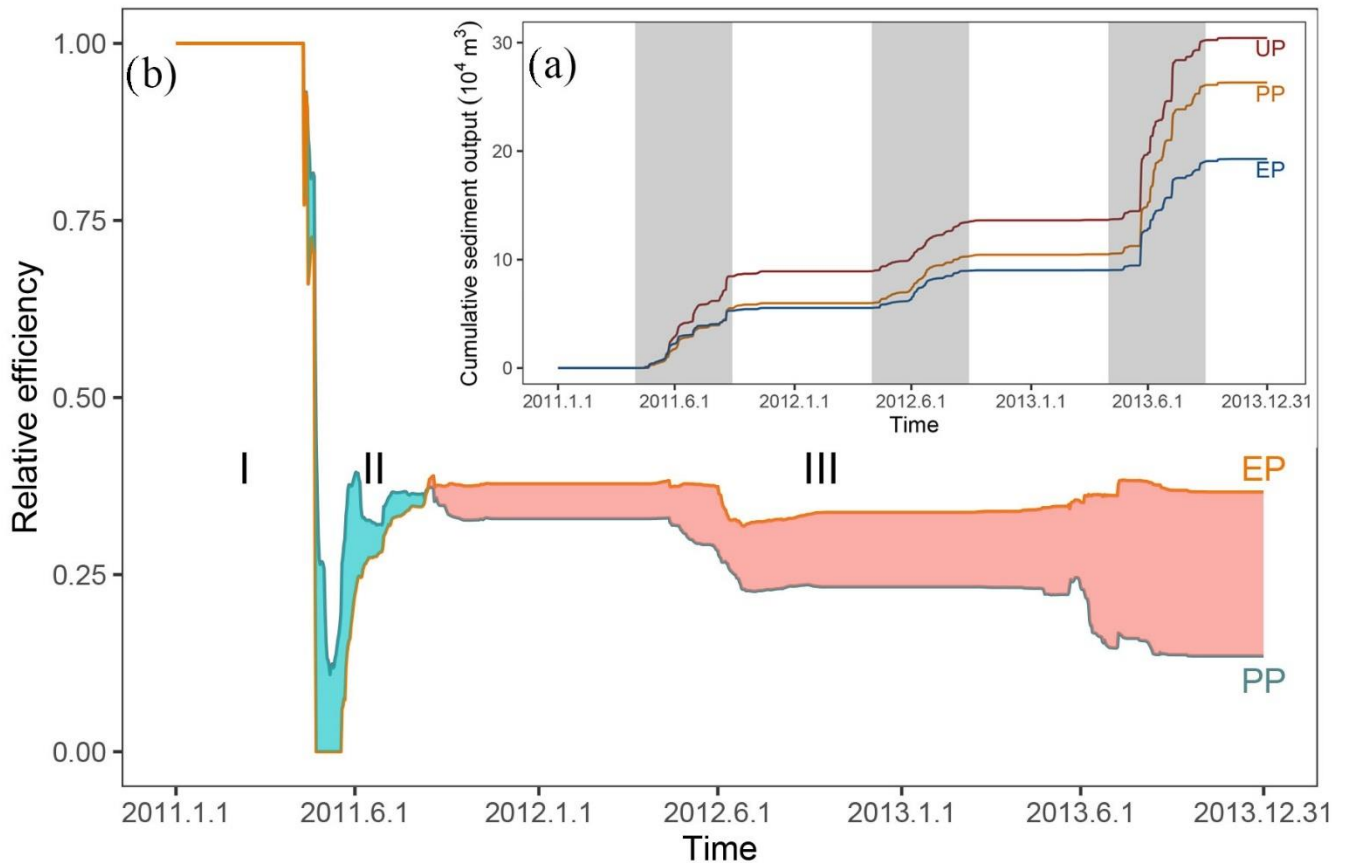
317 Compared with the Ca of the source area in Scenario UP, the value was increased by 138.1% in Scenario PP, which was
318 attributed to the dam1. Likewise, dam 2 in the transitional area reduced sediment loss effectively, which was reflected by a
319 52.5% increase in Ca . Furthermore, the mitigation measures in Scenario PP with vegetation revetment and levees in Scenario
320 EP worked better. The conservation ability in the source area increased by 161.9% due to the dam retainment and vegetation
321 revetment, and the levees helped increase by 3.49% in the deposit area.



322
 323 **Figure 9: The volumes of sediment and the conservation ability (*Ca*) in three areas for each scenario (S: source area; T: transitional**
 324 **area; D: deposit area).**

325 The cumulative sediment yield time series for each scenario and the relative efficiency of scenario UP and EP are present in
 326 Fig. 10b and Fig. 10a, respectively. The steep curve of output cumulative sediment means a significant increase in the deposi-
 327 tion. Three increasing stages are consistent with the rainfall intensity in three monsoons (May-September). The total sediment
 328 output in UP was the largest of $\sim 30.4 \times 10^4 \text{ m}^3$, followed by sediment yield in PP ($26.3 \times 10^4 \text{ m}^3$), and EP presented the least
 329 figure ($19.3 \times 10^4 \text{ m}^3$).

330 The relative efficiency over the period of controlling measures by human intervention in PP and EP (Fig. 10b) indicates three
 331 distinct stages. Stage I shows that the intervention measures in both scenarios prevented the sediment transport completely.
 332 Later stage II is a peculiar period when the effect of enhanced protective measures in EP was not as good as that in PP through
 333 repeated experiments. For stage III, the relative efficiency of the intervention measures in EP was greater than that in UP for a
 334 more prolonged stage III, which could achieve the long-term effect and stable conservation of solid material.



335

336 **Figure 10: (a) cumulative sediment yield over time (grey region highlighting three monsoons), (b) relative efficiency of scenario UP**
 337 **and EP compared with the UP (cyan shading represents when PP is more effective than EP and red shading represents the opposite)**

338 5. Discussion

339 5.1 Model calibration and uncertainty

340 The calibration and uncertainty are essential issues in the CAESAR-Lisflood (C-L) simulation of the geomorphic response to
 341 intervention measures (Yeh and Li, 2006). A preliminary calibration was carried out by replicating the geomorphic changes
 342 and water depth driven by an extreme rainfall event that occurred in 2018. The results (Fig. S3) demonstrated that the C-L
 343 successfully replicated the flash flood event using the initial conditions and model parameters. Actually, the calibration of the
 344 replicating ability of the geomorphic response to intervention measures was derived from a direct comparison between model
 345 results and direct measurements (Fig. 4 and Fig. 5). As a result, the simulated water discharge was more than the measured
 346 one but with the same order of magnitude. Moreover, the errors of erosion and deposition depth between simulation in Scenario
 347 PP and photographic evidence in three locations were at most 20%. The results suggest the robustness of the model settings
 348 and parameterisation.

349 The source of uncertainty is mainly from the model parameters and driving factors. Skinner et al. (2018b) provided a detailed
 350 sensitivity analysis of C-L, indicating that the sediment transport formula significantly influences a smaller catchment mod-
 351 elled by 10 m-grid cells. The sediment transport law, Wilcock and Crowe equations (Wilcock et al., 2003) have been proven
 352 suitable in the Xingping valley (Xie et al., 2018, 2022a, b; Li et al., 2020). Nevertheless, the empirical models of sediment
 353 transport would overpredict bedload transport rates in steep streams (gradients greater than 3%) (D'Agostino and Lenzi, 1999;
 354 Yager et al., 2012). Additionally, the driving factor, the input hourly rainfall data downscaled from the daily sequence is an
 355 unrealistic situation. Various sediment transport equations and downscaled hourly rainfall data need to be tested in C-L to
 356 determine the uncertainty further.

357 **5.2 Intervention effects**

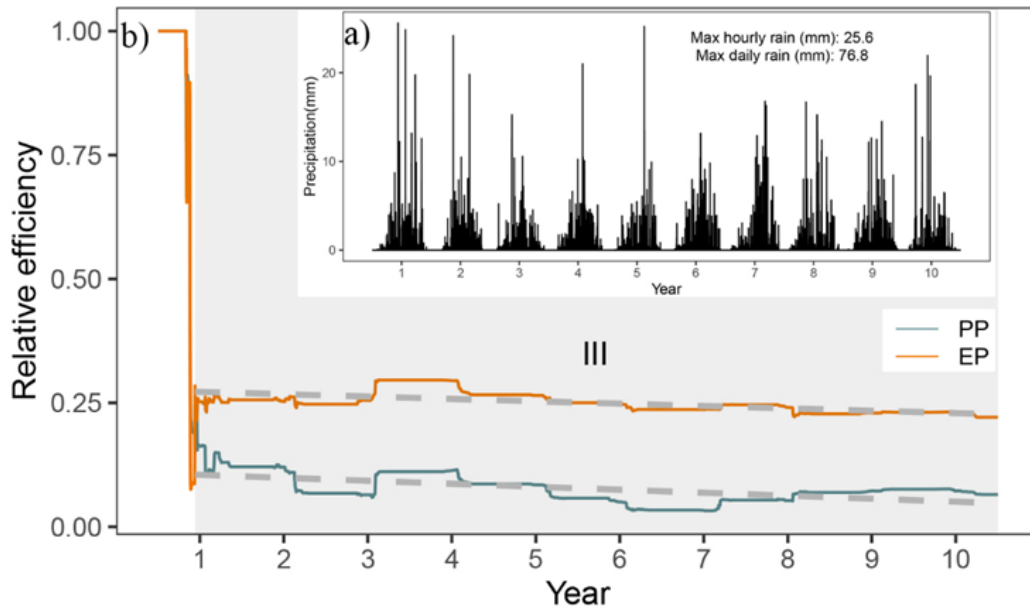
358 The more facilities equivalent to more comprehensive intervention systems in this study aim to control sediment delivery. The
359 C-L simulation of the geomorphic response to intervention measures suggests considerable influence on spatial modification
360 and sediment yield. The relative systematic measures lead to fewer total affected areas (7.9%-19.7%) and sediment yield
361 (16.7%~36.7%), which are suggested in overall evidence (see Fig. 6 and Fig. 10). Such catchment-scale performance disturbed
362 by the extreme event is in line with the viewpoints of other authors (Chen et al., 2023; Lan et al., 2020; Chen et al., 2015).

363 The mitigation measures considerably changed the soil conservation ability in three sub-regions, especially in the source area.
364 Herein, two main reasons why the erosion material is less in the source area than in another two sub-regions are inferred from
365 the interactions of loose material and topographic constraints. First, the abundant loose solid material formed by the strong
366 earthquake have stabilised overall since 2008's debris flow (details in Table S1). Second, the long and deep gullies are mainly
367 located in the transitional area (Yaogouli, Shicouzi, Yangjiashan) and deposit area (Qinggangping), which provide more sedi-
368 ment supply than the source area. As shown in Fig. S4, the movement of the material occurred mainly in the branch valleys in
369 the transitional and deposit area.

370 Moreover, comparing details and conservation ability with three scenarios stressed the unique role played by different inter-
371 vention measures. For example, the check dams are most effective in blocking sediment, and the vegetation revetments
372 strengthen the conservation ability. The synergy effect of soil conservation ability of checks dams and vegetable coverage is
373 created with an increase of more than two times. The levees are barriers with a discernable impact on sediment conservation
374 but with a special specific object-oriented protection.

375 The effectiveness of mitigation measures detected will decrease over time with a smaller downward trend. We supplement a
376 ten-year experiment to reveal the decreasing trend over a more extended period. We randomly selected one of the 50 repeated
377 rainfall datasets (year 2016-year 2025) downscaled by Li et al., 2020, which were generated from NEX-GDDP product (spatial
378 resolution: $0.25^{\circ} \times 0.25^{\circ}$, temporal resolution: daily) under RCP 4.5 emission scenario. The extracted rainfall sequence was then
379 input to C-L to simulate the effectiveness of three intervention scenarios. The result (Fig. 11) illustrates that stage III (the stable
380 stage that started on the 161st day, in which Scenario EP's intervention measures were more effective) lasted longer than stages
381 I and II, which were only at the start. The relative effectiveness in both PP and EP scenarios decreased gradually, while the
382 curve fell faster in PP than in EP.

383 The storage capacity of checking dams fades as the accumulation of sediment deposits, which necessarily lead to the gradual
384 decrease of intervention effectiveness. Additionally, the vegetation revetments remain operationally effective in reducing sed-
385 iment transport by stabilising topsoil over the period when the role of dam reservoirs gradually fails due to the lack of dredging
386 work. Therefore, the vegetation protection strategy is vital for “green development” to reduce sediment loss but requires further
387 efforts.



388

389 **Figure 11: Future rainfall input and relative efficiency of sediment intervention measures. (a) rainfall downscaled from stochastic**
 390 **future rain; (b) the relative efficiency changes over ten years (grey region highlighting stage III, and the grey dashed lines indicating**
 391 **the linear fitting curve).**

392 5.2 Limitations and applications

393 We built the dams and levees in our simulations by increasing the elevation in the expected location and assuming that it could
 394 not be eroded (see <https://sourceforge.net/projects/caesar-lisflood/>). This method proved experimentally feasible (Poepl et al.,
 395 2019; Gioia and Schiattarella, 2020). The rigid dam and levee body embedded in the model would not be broken or weakened
 396 over time so the simulation result could underestimate the geo-hazard risk. Considering the complexity of the geo-hazard
 397 mechanism, the abovementioned tools could not simulate the occurring process of geo-hazard chain links. They would ignore
 398 the fierce attack on the environment and facilities downstream. Some typical geo-hazard chains were focused on the specified
 399 event in a short time and recreated the hazard lifecycle using physical and mechanical models (Fan et al., 2020).

400 The methods applied in the study further demonstrate that C-L is an effective tool for understanding the short-medium term or
 401 long-term geomorphology changes (Ramirez et al., 2022; Li et al., 2020; Coulthard et al., 2012a) and observing the effective-
 402 ness of natural hazard interventions measures under different rainfall patterns. Our simulations indicate that the mitigation
 403 facilities in this study were effective, especially engineering efforts cooperating with vegetation revetments in the upstream
 404 area, which would help decision-makers to optimise the management strategies to control mountain disasters. Geotechnical
 405 engineering has disadvantages, even though it is a mature technology that identifies and fixes problems quickly (Cui and Lin,
 406 2013), such as the greater work and expense and the difficulty of maintenance. While the “green development”, the vegetation
 407 cover was effective in preventing erosion by strengthening topsoil and absorbing excess rainwater with its roots (Reichenbach
 408 et al., 2014; Stokes et al., 2014; Forbes and Broadhead, 2013; Mickovski et al., 2007). Alternatively, the methods could be
 409 used to study the tree planting patterns on different slopes.

410 6. Conclusions

411 In this study, the scenarios intervened by check dams, biological measures and artificial barriers are simulated using the C-L
 412 to outline the erosion and deposition area, measure the impacts of blocking sediment, and examine how the vegetable revet-
 413 ments helped to stabilise the slope. Four key findings are concluded. First, the engineering measures in controlling sediment
 414 transport are efficient, and the performance in protecting the fragile environment would be improved combined with other

415 intervention measures like vegetation revetment and artificial barriers. Second, the effectiveness of mitigation measures would
416 decrease over time. Third, the characteristics of sediment transport patterns changed considerably caused of the intervention
417 measures. The stabilising sediment ability in the source area increased by 161.9% with the additional effect of vegetation
418 revetments. At last, the present intervention measures are inadequate to reduce erosion and should be combined with dredging
419 work.

420 **Declaration of interest statement**

421 The authors declare that they have no known competing financial interests or personal relationships that could have appeared
422 to influence the work reported in this paper.

423 **Author contribution**

424 Di Wang: Conceptualisation, Methodology, Software, Writing-original draft preparation. Ming Wang Kai Liu and Jun Xie:
425 Supervision, Methodology, Writing- Reviewing and Editing, Validation.

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430

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