

Impact of large wildfires on PM10 levels and human mortality in Portugal

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Abstract.

Uncontrolled wildfires have a substantial impact on the environment, the economy and local populations. According to the European Forest Fire Information System (EFFIS), between the years 2000 and 2013 wildfires burnt about 170,000-740, up to 740,000 ha of land annually on the south of Europe (Portugal, Spain, Italy, Greece and France). Although most southern

5 European countries have been impacted by wildfires in the last decades, Portugal was the most affected, having, being Portugal the country with the highest percentage of burned area comparing to its whole territory. For this reason, it deserves a closer attention per km². However, there is still a lack of knowledge regarding the impacts of the wildfire-related pollutants on the mortality of the country's population. All wildfires occurring during the fire seasons season (June-July-August-September) from 2001 and 2016 were identified and those with a burned area above 1000 ha-1,000 ha (large fires) were considered for
10 the study. During the studied period (2001-2016), more than 2 million ha of forest (929,766 ha from June to September alone) were burned in mainland Portugal. Although large fires only represent less than 1% of the number of total fires, in terms of burned area their contribution is of 46% (53% from June to September). To assess the spatial impact of the wildfires, these burned areas in each region of Portugal were correlated with PM10 concentrations measured at nearby background air quality monitoring stations, provided by the Portuguese Environment Agency (APA). Associations between PM10 and all-
15 cause (excluding injuries, poisoning and external causes) and cause-specific mortality (circulatory and respiratory), provided by Statistics Portugal, were studied for the affected populations, using Poisson regression models. During the studied period (2001-2016), more than 2 million ha of forest were burned in mainland Portugal and the 48% of wildfires occurred were large fires. A significant A significant positive correlation between burned area and PM10 have been was found in some NUTS III (regions) on regions of Portugal, as well as a significant correlation between burned area and mortality. North association
20 between PM10 concentrations and mortality, being these apparently related to large wildfires in some of the regions. The north,

centre and inland of Portugal are the most affected areas. The high temperatures and long episodes of drought expected ~~on-in~~ the future will increase the probabilities of extreme events and therefore, the occurrence of wildfires.

1 Introduction

~~The existence of wildfires constitutes~~ Wildfires have a considerable impact on the environment and humans ~~living in numerous regions~~ worldwide. Climate change has lately been identified as a very important variable ~~to be considered~~ in this matter ~~and global warming scenarios are forecasting~~ (Gillett et al., 2004) since ~~the future projections suggest~~ an increase of the number and intensity of wildfires ~~during the next years~~ (Bowman et al., 2017) ~~of droughts, heat waves and dry spells~~ (Turco et al., 2019). Global warming will produce changes in temperature and precipitation patterns ~~which will increase the~~ ~~leading to a higher~~ prevalence and severity of wildfires (Settele et al., 2015), (Settele et al., 2015; Bowman et al., 2017) and consequently impacting future air quality (Schär et al., 2004). In fact, ~~an increase on the number of droughts, heat waves and dry spells is suggested by climate change projections, which this~~ could not only extend the burnt area in chronically impacted areas (Cisca et al., 2014), but also affect new ones (Gillett et al., 2004), as was the case of, ~~like~~ Sweden in the summer of 2018 (Lidskog et al., 2019). ~~One of the areas of the world most fustigated by wildfires is the Mediterranean basin (Portugal being the most impacted country), which needs to be studied carefully to address the concerns of local populations.~~

According to the 2016 European Forest Fire Information System (EFFIS) report (San-Miguel-Ayanz et al., 2017), the south of Europe (Portugal, Spain, France, Italy and Greece) is the area most affected by wildfires since 1980 until today, considering Europe, Middle East and North Africa. In the last decades, Portugal was by far the country with the largest burned area, almost 50% between the southern European countries (Parente et al., 2018). Although there has been a slight decreasing trend in the burnt area in this region since 2000 after an increasing period in the previous 20 years (European Environment, Agency, <https://www.eea.europa.eu/data-and-maps/indicators/forest-fire-danger-3/assessment>), recent extreme events like the 2017 fires in Portugal and the 2018 fires in Greece which ~~even~~ resulted in a severe loss of human lives are confirming the worst projections. ~~In fact, a recent study by Turco et al. (2019) showed a relationship between drought and the occurrence of wildfires and suggested an increase of both due to future climate change. But already some years before, the PESETA (Projection of Economic impacts of climate change in Sectors of the EU based on bottom-up Analysis) study estimated an increase of the burnt area in southern Europe in the future (Cisca et al., 2014).~~

Uncontrolled wildfires emit numerous pollutants derived from the incomplete combustion of biomass fuel, which cause damage to human health, particularly the ~~respiratory system~~ ~~cardiovascular and respiratory systems~~ (World Health Organization, 2010). Examples include particulate matter (PM), carbon monoxide, methane, nitrous oxide, nitrogen oxides, volatile organic compounds (VOCs), and other secondary pollutants (Cascio, 2018) that are released mainly into the atmosphere but can be transported to many other environmental compartments. Moreover, they can affect the physicochemical properties of the atmosphere, as for instance the interaction of PM with solar radiation which can prompt a modification of the temperature depending on the characteristics of the aerosol (Trentmann et al., 2005). Consequently, some of these chemicals are regulated by the European Directive 2008/50/EC of 21 May 2008 of the European Parliament and of the Council on Ambient Air Quality

and Cleaner Air for Europe, which establishes threshold values for a safe air quality. But although wildfire emissions are a crucial parameter for the local air quality (Knorr et al., 2016), where in some cases there are already chronically-exposed populations due to the frequency and dimension of the events, they are not contained by political borders and can also affect areas far from the ignition points due to the atmospheric transport of the pollutant plumes. A number of studies (e.g. Lin et al. (2012); Im et al. (2018); Liang et al. (2018); Augusto et al. (2020), among others) report the influence of natural and anthropogenic emissions on air quality composition across different countries, especially PM and tropospheric O₃. For wildfires it is also important to take into account some factors which influence the plume dispersion, such as the duration and space evolution of the fire event and the meteorological conditions associated (Lazaridis et al., 2008). An increase of cardiovascular and respiratory morbidity and mortality are some of the impact these contaminants can have on humans (Johnston et al., 2012; Tarín-Carrasco et al., 2019). For instance, there is a strong evidence of the relationship between PM in general and mortality, especially from cardiovascular diseases, for both long-term and short-term exposure (Anderson et al., 2012). Although some studies corroborate the existence of a link between the exposure to wildfire-related air pollutants and hospital admissions, visits to emergency clinics or even respiratory morbidity (Liu et al., 2015; Reid et al., 2016), the impacts on human health are difficult to quantify and the real effects still poorly known.

Regarding Concerning PM, a recent study focusing on 10 southern European cities revealed that cardiovascular and respiratory mortality associated to PM10 (particles with aerodynamic diameter below 10 μm) was higher on days affected by wildfires' smoke than in smoke-free days (Faustini et al., 2015). The authors also found that PM10 from forest fires increased mortality more than PM10 from other sources. So, the estimation of mortality due to exposure to wildfire-generated pollutants is key to manage health resources and the necessary public funds towards prevention and remediation, setting up appropriate policies and protocols (Rappold et al., 2012).

The two main factors to take into account for the wildfire's effects are the location and, most importantly, the size of the fire event (characterised by the respective burnt area). When the wildfire occurs close to a large conurbation, the population exposed is higher. But as Analitis et al. (2012) showed in their study, small fires do not seem to have an effect on mortality, whereas medium and large episodes (with burnt areas >1000 ha) have a significant impact on human health, which increases with the size of the fire. Aiming to enhance the knowledge on the effects of wildfires on human health, this study describes the pattern of wildfires in Portugal for 16 years (2001-2016) and assesses the impact of those events on the country's population mortality during the fire season (June, July, August and September). In this work, the focus is placed on indirect effects of pollutants emitted by wildfires, namely assessing the influence of wildfire-generated PM10 on the Portuguese population mortality. The relationship between the size of the wildfire (characterised by the respective burned area) burned area of large wildfires and PM10 and this same pollutant and mortality has been was studied. The Nomenclature of Territorial Units for Statistics (NUTS) level 3 (NUTS III) geographical division has been used to be able to compare the effects of the fires in different parts of the country. Finally, monthly deaths due to all-cause (excluding injuries, poisoning and external causes) and cause-specific mortality (cardiovascular and respiratory) for all ages for each NUTS III has been studied. These causes have been selected due to their well-known connection with air pollution.

2 Methodology and data

In this study, the effects of short-term ~~pollutants exposure~~ exposure to pollutants due to wildfires on human mortality ~~are~~ were quantified. The forest fire pollutant emissions were estimated for the period 2001-2016 during the summer months (June-July-August-September) in ~~Portugal mainland~~ mainland Portugal (23 NUTS III and more than 10 million people). ~~For this In Portugal, large forest fires usually occur during the months of June, July, August, and September, which correspond to the time of the year with the highest temperatures and driest conditions. By focusing our study only on these 4 months, we can have enough data to perform a valid statistical treatment, while avoiding a strong influence of PM10 from other sources in colder months (such as home heating or traffic). At the same time, we do not include in the analysis the deaths due to, for instance, cold and flu that could become confounding factors.~~

~~For the quantification, two steps have been followed:~~

~~were followed.~~ First, an assessment of the incidence, patterns and variations of burned area on a large time frame and spatially integrated by NUTS III was done on the levels of air pollutants. PM10 and burned area ~~has been~~ were correlated through linear regression, while the mortality data and PM10 ~~was~~ were correlated with Poisson regression. Data was processed and ordered by NUTS and by month and year. Finally, the correlation between the pollutants emitted by forest fires, the wildfires burned area and the different causes of mortality during the period 2001-2016 for the summer months was studied. The study is focused ~~in~~ on PM since it is one of the main pollutants emitted by wildfires, which can increase PM concentrations up to 50% and more (Lazaridis et al., 2008). Moreover, there is a clear relation with several effects on human health (including mortality), in particular with respiratory and circulatory diseases (~~?Reid et al., 2016; Liang et al., 2018~~ (Kollanus et al., 2016; Reid et al., 2016; Liang et al., 2018)). There was not enough PM2.5 data collected from the Portuguese air quality management network to establish a correlation (only 20 stations measure PM2.5 in the mainland). For all these reasons this study ~~is focused~~ focuses on PM10.

2.1 Target area

With 89,015 km² (9.11 Mha) mainland Portugal accounts for over 96% of the country's area and hosts over 10 million inhabitants in the west Iberian Peninsula (southwestern Europe). With the largest urban areas along the west Atlantic coast, particularly around the capital ~~Lisbon~~ more to the south (~~Lisbon~~) -more to the south- and the second largest city (Porto) in the north (see Figure 1, left), the country has most of its mountain ranges in the north, reaching ~~1993 m~~ 1,993 metres above sea level in Serra da Estrela. Although showing a Mediterranean climate, this topographic display leads to various climate patterns along the country, with increasing temperature and decreasing rainfall from northwest to southeast (Moreira et al., 2011; Oliveira et al., 2017). In terms of land cover, Figure 1, left shows a predominance of agriculture by 2015 (over 50% and mainly in the south), followed by forests and shrublands, which comprise 43% of the territory (mainly in the north and southwest). This ~~allied~~ combined with high temperatures in the summer months ~~represent~~ represents a potential fire hazard, which unfortunately has been often ~~verified~~ proved almost every summer for many years.

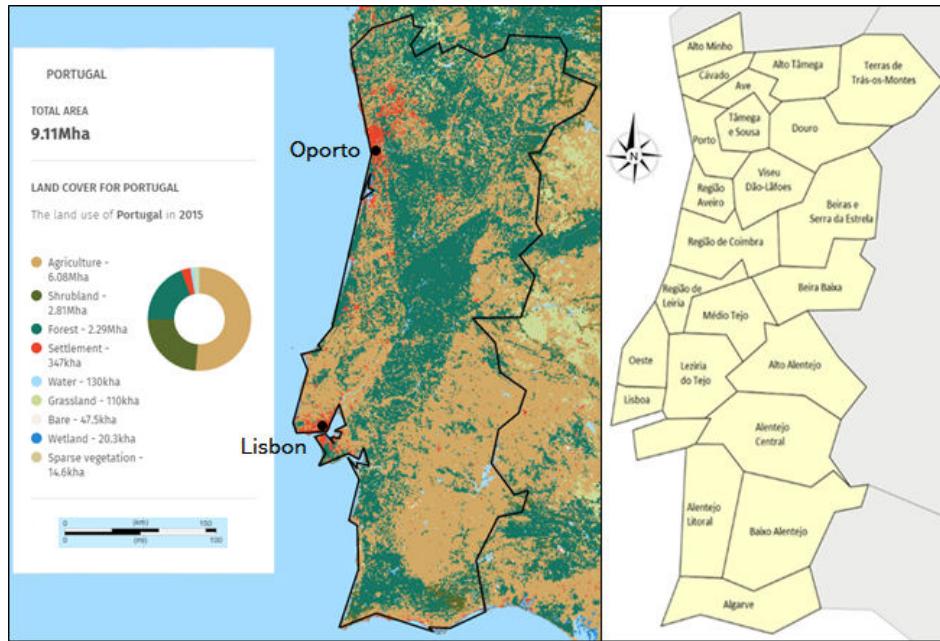


Figure 1. (Left) Land cover in mainland Portugal in 2015 (Global Forest Watch, <https://www.globalforestwatch.org>; (Right) Mainland Portugal NUTS III regions as included in this contribution.

120 2.2 Datasets

2.2.1 NUTS III boundary data

The target domain was divided by NUTS (Nomenclature of Territorial Units for Statistics) level 3 (Figure 1, right) for Portugal mainland. NUTS is a geocode standard for referencing the subdivisions of countries for statistical purposes developed by the European Union, and are divided in three levels (I, II, III) which are established by each EU member country.

125 The boundaries of the NUTS-III files NUTS III from mainland Portugal (in total, 23) at a 1:60 million scale were retrieved from the Eurostat web page (Eurostat, 2019) and treated with QGIS3 software. The downloaded data are for the 14 March 2019 version at a 1:60 million scale.

2.2.2 Wildfires data

The wildfire data, collected in for the period from 2001 and 2016, was obtained from the Portuguese Institute for Nature 130 Conservation and Forests (<https://www.icnf.pt/>). For this study, forest fires occurring in the months of only forest fires were considered, and from them, only those with more than 1,000 ha of total burned area (which we denominated *large fires*) were selected. In total, there were 323 events under that category (less than 1% of the number of total fires), which were responsible for 46% of the total burned area. Table 1 shows the yearly variability during the studied period of the number of the occurrences,

total burned area or the contribution of large fires to the burned area in mainland Portugal. Although other studies have shown a relationship with high temperatures and drought periods (Turco et al., 2019), the data in Table 1 suggests that it is not possible to perceive a yearly pattern of wildfires in the country. For instance, in 2008 no large fires occurred, whereas 2003 accounted for the highest number of occurrences (81), which were responsible for 80% of the total burned area in that year. But the latter contribution was as low as 12% in 2011 and had a mean percentage for the whole period of 34%.

Table 1. Number of wildfires and burned area (BA) by year for the period 2001-2016 in mainland Portugal. From left to right: number of occurrences when the burned area is larger than 1000 ha; sum of the burned area for fires larger than 1000 ha; total burned area caused by all fires; percentage of burned area caused by large fires; and index between burned area and the number of occurrences.

| Year | Occ. (N) with BA > 1000 ha | BA > 1000 ha | Total BA | % BA > 1000 ha | BA/Occ. (ha/N) |
|------|----------------------------|--------------|----------|----------------|----------------|
| 2016 | 22 | 85166 | 138884 | 61 | 3871 |
| 2015 | 8 | 16629 | 63227 | 26 | 2078 |
| 2014 | 3 | 5560 | 19771 | 28 | 1853 |
| 2013 | 26 | 66633 | 152181 | 44 | 2563 |
| 2012 | 11 | 41884 | 108965 | 38 | 3808 |
| 2011 | 6 | 8694 | 72006 | 12 | 1449 |
| 2010 | 25 | 54901 | 131753 | 42 | 2196 |
| 2009 | 9 | 18018 | 86478 | 21 | 2002 |
| 2008 | 0 | 0 | 16662 | 0 | - |
| 2007 | 2 | 4623 | 32595 | 14 | 2312 |
| 2006 | 7 | 19008 | 75513 | 25 | 2715 |
| 2005 | 61 | 167133 | 338262 | 49 | 2740 |
| 2004 | 24 | 56088 | 129715 | 43 | 2337 |
| 2003 | 81 | 338486 | 425742 | 80 | 4179 |
| 2002 | 17 | 29414 | 124408 | 24 | 1730 |
| 2001 | 21 | 32509 | 112166 | 29 | 1548 |

Considering only June, July, August and September 2001-2016 (the months with highest temperatures and drier conditions when more than 65% of fires happened) with more than 1000 ha of total burned area were selected and considered large fires. In total, there were 331 events under that category. This 86% of the total fires and 311 of the 323 large fires -96% occurred-, this data was divided by month and year and the respective monthly and yearly sums were considered for each NUTS III level region. Ave, Alto Tâmega, Tâmega e Sousa, Oeste, All NUTS III had at least one large fire during the study period. In terms of burned area 929,766 ha of forest were lost in mainland Portugal from June to September (2001 to 2016), with about 53% due to large fires (>1,000 ha). Figure 2 represents the number of large fires and the burned area they were responsible for, by NUTS III.

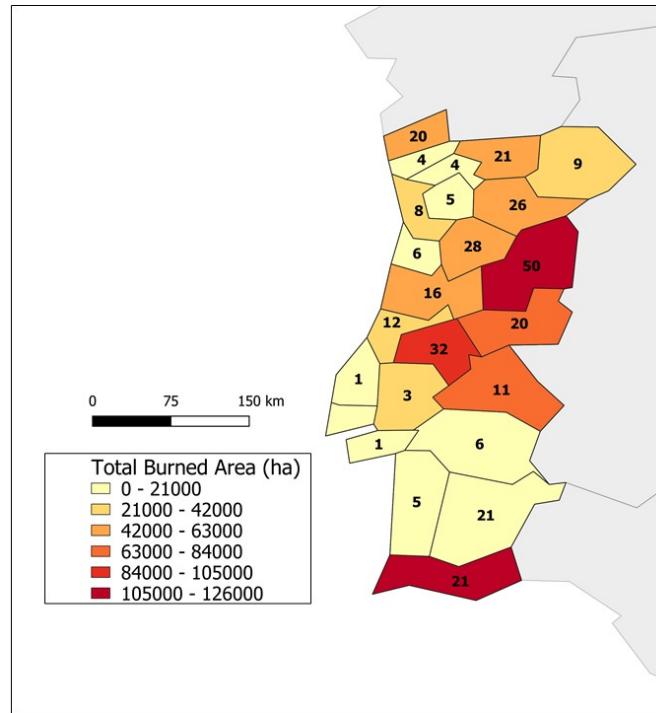


Figure 2. Burned area caused by large fires and number of fires ($>1,000$ ha) per NUTS III from June to September in the period 2001-2016.

The north and centre of Portugal present the most extensive forest cover in the country (Nunes et al., 2019), particularly abundant in pine and eucalyptus trees, two highly combustible species that have been associated with extreme wildfire events (Maia et al., 2014). Consequently, both areas show the highest number of large fires and respective burned area (being Beiras e Serra da Estrela and Médio Tejo ~~and Alentejo Litoral~~ the most affected NUTS III), but with also Alto Alentejo and Algarve (more to the south, see Figure 1) among the NUTS III with a higher incidence. Additionally, dense Mediterranean forests over hard-to-reach mountains can also be found in these areas, which combined enhance the difficulty of the firefighting efforts. Algarve, despite being located in the south coast, also has some mountains with forests, surrounded by a considerably dry and arid terrain, especially in the summer (Nunes et al., 2019), leading to a burned area of 112,764 ha, the second highest at the NUTS III level. Beiras e Serra da Estrela is the region which presented the largest burned area in the summer months from 2001 to 2016, (almost 117,000 ha) and the highest number of large fires (50). On the other hand, Oeste and A.M. Lisboa (mainly non-forested areas) are the NUTS III ~~where large fires during the study period were not found~~ with a smaller number of large fires, only one during the target timeframe. More detailed information can be found in Supplementary Material (Table SM1).

2.2.3 Pollution data

160 The information available on the levels of pollutants was obtained from the Portuguese Environment Agency air quality network (<https://qualar.apambiente.pt/qualar/index.php>), established to monitor the concentrations of pollutants according to the

European Legislation requirements (European Directive 2008/50/EC of 21 May 2008). The ~~network~~-location of the air quality stations is irregularly scattered throughout the country, with a stronger presence in the most populated areas (Figure 3). The isolation of pollutant emissions due to burnt biomass is quite complicated as it depends on parameters, such as vegetation type, the weather conditions on the ~~burnt moment~~ moment the fire is taking place or the contribution of other sources, among others. For this reason, in this study, background stations (specifically urban, suburban and rural background) were selected, so that the direct impact of other urban and industrial sources such as road traffic, building heating and manufacturing combustions was avoided.

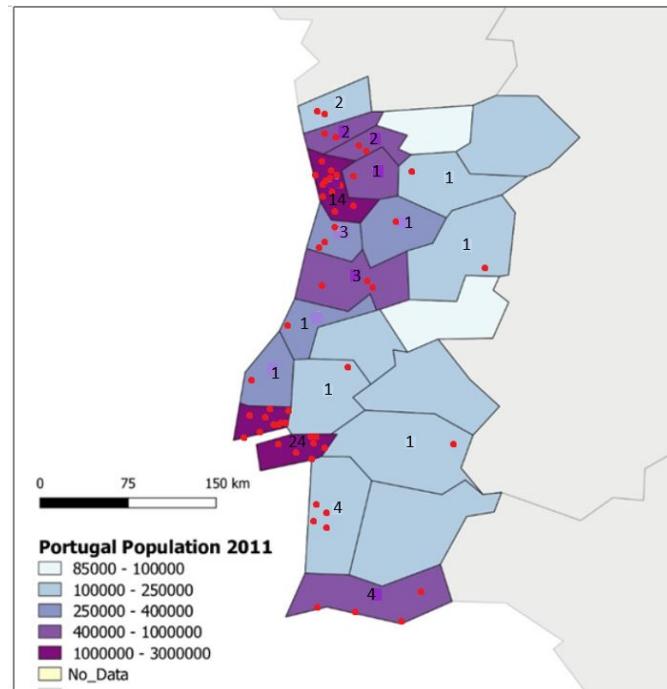


Figure 3. Population of each NUTS III according to the 2011 Census (<https://www.ine.pt>) and respective number of monitoring stations for PM10. Red dots indicate the location of the air quality monitoring stations used in this study.

Considering all the pollutants measured on the background stations, PM-PM10 was the one with a potentially higher link to forest fires. ~~And although~~ Although some stations also measured PM2.5, the coverage ~~in this case~~ was insufficient to draw any significant correlations, ~~so~~. The main anthropogenic sources of PM10 ~~was chosen in the end~~ include road traffic, industrial activities, and home heating. In this study, to minimize the influence of non-wildfire causes for the PM10 concentrations, we selected only background stations (encompassing urban and semi-urban ones, which are located within urban areas but with minimum influence of road traffic; and rural stations). Therefore, urban stations with road traffic influence and stations close to industrial complexes were not selected. The influence of home heating was already minimized by selecting the summer period as our target timeframe.

As done before for the wildfires data, also here the time range considered was from 2001 to 2016 and only the months of June to Septemberwere considered, with monthly means used for the correlations. Concentrations of PM10 were obtained for mainland Portugal in all types of background stations (a total of 91 which cover 17 NUTS III, as shown in Figure 2-3). Given the uneven coverage of the target domain, most stations are located in the metropolitan areas of Oporto (14 stations) and Lisbon (24 stations) and in the rest of the coastal areas, where the higher population (NUTS III commonly above 250,000 inhabitants, Figure 23) demands a tighter control of the air quality, but where, in turn, not a lot of large wildfires occur due to the urbanized land use. For the NUTS III with more than one station a mean between all the PM10 concentrations in each NUTS was done.

Figure 4 shows the mean concentration of PM10 by NUTS III from June to September in the period of 2001-2016 in the stations available in mainland Portugal. The highest mean concentrations of PM10 during the period 2001-2016 (June to September only) were observed in Oporto, with $31 \mu\text{g m}^{-3}$; followed by Lisbon, Alentejo Central and Ave, with levels ranging from 26 to $29 \mu\text{g m}^{-3}$ (Figure 4). Conversely, the NUTS III which present the lowest mean values, between $14-17 \mu\text{g m}^{-3}$, are Oeste, Alto-Minho and Viseu Dão-Lafões. Despite these values, no NUTS in Portugal exceed the threshold value of PM10 ($40 \mu\text{g m}^{-3}$ per year) established by the European Directive 2008/50/EC.

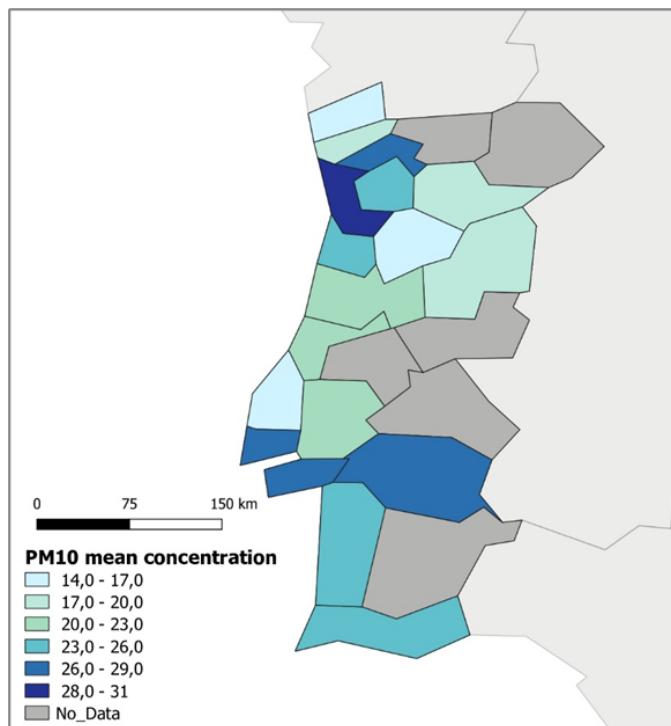


Figure 4. Population Mean concentration of each PM10 ($\mu\text{g m}^{-3}$) by NUTS III according from June to September in the 2011 Census and respective number period of monitoring stations for PM10 2001-2016 in mainland Portugal.

190 **2.2.4 Mortality data**

Mortality data covering the period from 2001 to 2016 was obtained from Statistics Portugal (<https://www.ine.pt>). Monthly death counts due to all-cause (International Classification of Diseases (ICD-10), codes A00-R99) excluding injuries, poisoning and external causes; and cause-specific mortality: cardiovascular (codes I00-I99) and respiratory (J00-J99) were collected for each NUTS III region of Portugal, comprising all-age residents. These mortality causes were selected since they have been reported previously in literature as important in their connection with air pollution (Hoek et al., 2013; Liu et al., 2015; Münzel et al., 2018) (Hoek et al., 2013; Liu et al., 2015; Kollanus et al., 2016; Münzel et al., 2018), in particular with particulate matter (PM). Other relevant mortality causes, such as Chronic Obstructive Pulmonary Disease (COPD, codes J40-J45) and asthma (ICD-10, code J47), were also considered, ~~however the reduced number of deaths due to these diseases in the study period prevented the establishment of correlations~~. However, since in many months and NUTS III in the target time period there were no deaths for COPD and asthma, it was not possible to obtain a data series large enough to correlate with PM10 and wildfires series.

2.3 Statistical analysis

2.3.1 Correlations between PM10 and burned area

~~Correlations~~ The correlations between PM10 and the total burned area ~~per month by large fires per month for each NUTS III~~ were estimated using Pearson correlations coefficients (Pearson and Galton, 1895). ~~Pearson correlation is The Pearson approach~~, used to correlate two continuous variables having a normal distribution, ~~while Poisson coefficients are used to correlate a count variable with a continuous variable. These methodologies are widely used in studies covering the topic of health impacts of~~ is widely found in studies of air pollution (e.g. Islam and Chowdhury (2017); Pallarés et al. (2019); Rahman et al. (2019); Pallarés et al. (2019); Rahman et al. (2019); Rovira et al. (2020); among many others). Results were considered statistically significant if the p-value was $p < 0.05$. Correlations were performed using the detrended data series of burned area and PM10 in order to remove the strong seasonal cycle of these variables and avoid spurious correlations. The detrending method follows Tarín-Carrasco et al. (2019), using the first-time difference time series.

2.3.2 Associations between burnt area, PM10 and mortality

The associations of monthly average PM10 levels, and the ~~size of the occurrence of large~~ wildfires (burnt area ~~>1000 ha and burnt area <1000 ha~~ 1,000 ha), with mean monthly mortalities (all-cause, respiratory and cardiovascular causes) were studied for the months ~~of~~ June, July, August and September for the period between 2001 and 2016. The ~~effect estimates estimates of the effects~~ were obtained for each NUTS III region using ~~Pearson regression models~~ Poisson regression models (Faustini et al., 2015; Islam and Chowdhury, 2017). Poisson coefficients can correlate a count variable (such as the number of deaths) with a continuous variable. The results were expressed as the Relative Risk (RR) of all-cause, cardiovascular and respiratory mortalities with a 95% confidence interval (95% CI). All regression models were performed using IBM SPSS Statistics 25.0 software.

3 Results and discussion

The results obtained in the study are presented as follows. First, the description of the situation in Portugal in terms of geographical distribution of the burned area is presented. Then, a summary of the PM10 concentrations during the summer months of the years 2001–2016 is presented. Finally, the correlations between burned area and PM10 and the potential associations of wildfire-derived PM10 and all-cause mortality is presented in this section.

3.1 Spatio-temporal patterns of wildfires

3.0.1 Burned area

From 2001 to 2016 period more than 2 million ha of forest were burned in mainland Portugal. Around 48% due to large fires

3.1 Relationship between burned area and particulate matter

For the correlation between the burned area from large fires and PM10, a significant positive correlation was found for 7 (out of 13 with available data) of the studied NUTS III, represented by the dotted areas in the map of Figure 5. For Oeste, Região de Leiria, Beira Baixa, Médio Tejo, Cávado, Ave, Terras de Trás-os-Montes and Alto Tâmega and the four Alentejo NUTS (>1000 ha). During this period, the wildfires occurred on different areas in Portugal, as shown in Figure 3.

Burned area and number of fires (>1000 ha) per NUTS III from June to September in the period 2001–2016.

The north, centre and inland of Portugal are the areas with the highest number of wildfires and the highest burned area, being Beiras e Serra da Estrela and Beira Baixa. Also Lezíria do Tejo (in Alentejo) and Algarve (in the south) (Figure 1) can be found among the most affected areas (in number of wildfires and burned area). As seen in Figure 3, the north and centre of Portugal is where the most extensive forests in the country (Nunes et al., 2019), particularly abundant in pine and eucalyptus trees, two species that have been associated with extreme wildfire events (Maia et al., 2014). Additionally, dense Mediterranean forests over hard-to-reach mountains can also be found in these areas, which combined enhance the difficulty of the firefighting efforts. The Algarve, which despite being located in the south coast, also has some mountains with forests, surrounded by a considerably dry and arid terrain, especially in the summer (Nunes et al., 2019). Beira Baixa is the region which presented the most burned area during the period 2001–2016, with almost 124,000 ha in total. But Beira e Serra da Estrela is the area where the highest number of wildfires was found, 47 in total. On the other hand, Lezíria do Tejo is the NUTS III region most affected by wildfires, considering the number of large fires and burned area together. Oeste and Área Metropolitana de Lisboa are the areas with a smaller number of large fires, only one during the target timeframe (since these are mainly non-forested areas). Table 1 presents an overview of the burned area and occurrences of large fires by NUTS III areas in mainland Portugal.

Total burned area and occurrences of *large fires* by NUTS III areas in mainland Portugal in the period 2001–2016 in the months of June, July, August and September (total of 64 target summer months). NUTS Months with large fires (N) Number of large fires (N) Total burned area (ha) **North** Alto Minho 7 14 53918 Cávado 3 4 7731 Ave 4 4 5410 **Alto Alentejo**, **Alentejo Central**, **Alentejo Litoral** and **Baixo Alentejo**), there was not enough pollutant and/or burned area data enough to establish

statistical relationships. The correlations are strongest for Cávado, Ave, Tâmega 10-21 48415 Terras de Trás-os-Montes 7-10 28050A. M. Porto 5-8 40840 Tâmega e Sousa 4-4 5706 Douro 14-27 48907 Centre Aveiro 3-6 17481 e Sousa, Região de Aveiro and Viseu Dão-Lafões 11-31 60744 Coimbra 8-16 45673 Beiras e Serra da Estrela 17-49 115503 Leiria 6-13 28347 Médio Tejo 11-31 96947 Beira Baixa 7-14 59032 Oeste 1-1 1700A. M. Lisboa 1-1 2756 Alentejo Lezíria do Tejo 2-3 24404 Alto Alentejo 4-12 70657 Alentejo Central 3-6 1970 Alentejo Litoral 3-5 16176 Baixo Alentejo 2-2 9240 Algarve 9-19 107273

On the other hand, Table 2 shows the yearly variability during the studied period of the number of the occurrences and burned area Dão-Lafões, with correlation coefficients above 0.75 at a confidence level of 0.95, followed by Alto Tâmega and Beiras e Serra da Estrela, between 0.5 and 0.74. As expected, all these areas are in the north and center of mainland Portugal, in line with the denser forest cover. In 2008 no wildfires over 1,000 ha of burned area occurred, whereas 2003 accounted for 81 of occurrences of large fires, which were responsible for 80% of the total burned area in that year. The data in Table 2 suggests that it is not possible to perceive a yearly pattern of wildfires in Portugal regarding the occurrences, burned area or the contribution of large fires to the burned area, but other studies have shown a relationship with high temperatures and drought periods (Tureo et al., 2019).

Number of wildfires and burned area (BA) by year for the period 2001-2016. From left to right: number of occurrences when the burned area is larger than 1000 ha; sum of burned area for fires larger than 1000 ha; percentage of burned area caused by large fires; and index between burned area and the number of occurrences. **Year** **Oee. (N)** **with BA>1000ha** **BA>1000ha** **Total BA** **%BA>1000ha** **BA/Oee. (ha/N)** 2016 2285166160458533871 2015 8168866497826211120143645121114312150201326713911566884627 2002 173078612973124181120012134856116706301660

According to the 2016 EFFIS report, the south of Europe (Portugal, Spain, France, Italy and Greece) is the area most affected by wildfires since 1980 until today. In the last decades, Portugal was by far the country with the largest burned area, almost 50% between the southern European countries (Parente et al., 2018). But not only the south of Europe seems to be affected by wildfires, as areas in north of Europe which never had relevant wildfires are now suffering from these extreme episodes, as was the case of Sweden in the summer of 2018 (Lidskog et al., 2019).

3.1.1 Particulate matter (PM10)

Regarding particulate matter, air quality monitoring stations are unequal spatial distributed. As mentioned previously, most of them are located near the coast, particularly in the in metropolitan areas of Lisbon and Oporto, the two most densely populated in the country, and only a few monitoring stations can be found in inland. The mean highest concentrations of Finally, in Alto Minho, A.M. Porto, Douro, Região de Coimbra, Região de Leiria and Algarve no significant correlations were found. The limited number of stations in those areas (which means fewer data to correlate), their location (closer or farther from the large wildfire spots) and uneven distribution, and the contribution of other sources for the PM10 during the period 2001-2016 (June to September only) were observed in Oporto, with $31 \mu\text{m m}^{-3}$; followed by Lisbon, Alentejo Central and Ave, with levels ranging from 26 to $29 \mu\text{m m}^{-3}$ (Figure 4) Conversely, the NUTS III which present the lowest mean values, between $14-17 \mu\text{m m}^{-3}$, are Oeste, Alto-Minho and Viseu Dão-Lafões. Despite these values, no NUTS in Portugal exceeded the threshold value of PM10 ($40 \mu\text{m m}^{-3}$ per year) established by the European Directive 2008/50/EC.

3.2 Relationship between burned area and particulate matter

Regarding correlation between burned area and PM10, a significant positive correlation was found for most of the studied NUTS III, represented by black dots in the map of Figure 5. The correlation is weaker for Alto Minho, Cávado and Algarve but 290 the rest of studied NUTS present a strong correlation between both variables, peaking in Lisbon and Leiria with more than 0.98 at a confidence level of 0.95. Obviously, the levels can be some of the explanations. The location of the air monitoring stations may play a key role in these correlations, especially when they are scarcer, but for these NUTS this is a good indication where the influence of wildfires on the emissions of PM10 is likely to be stronger. In fact, some authors have reported a contribution 295 of wood burning to the PM10 load even in urban environments, where the presence of other PM sources tends to be higher (Fuller et al., 2014; Perrino et al., 2019).

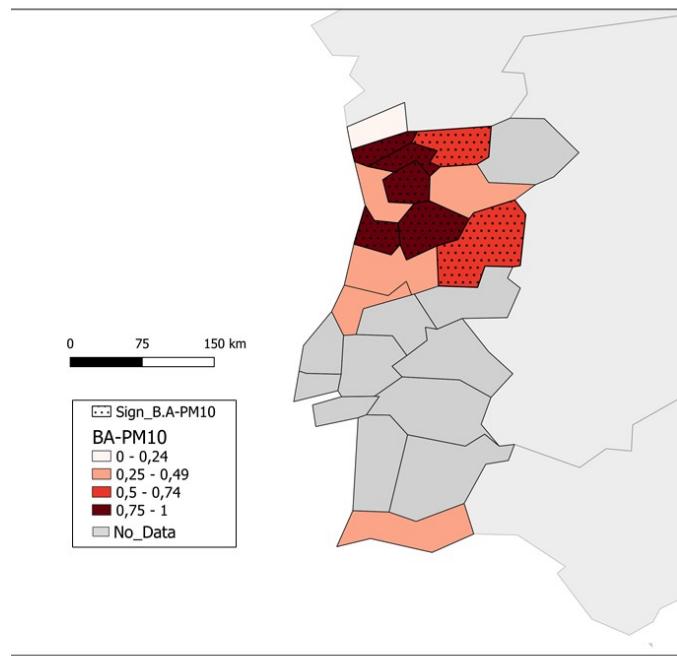


Figure 5. Significance of Pearson correlations between burned area and PM10 for each NUTS III in the period of 2001-2016 (from June to September; dots represent significant correlations at 95% confidence).

3.2 Impact of wildfires on **all-cause** mortality

3.2.1 Mortality overview

The mortality counts for the period 2001-2016 (for the months June to September, 64 months) in mainland Portugal are presented in Table 2, for each NUTS III region and all-cause, cardiovascular and respiratory-related deaths. Results show that

300 almost 30% of all-cause and cardiovascular mortality occur during the extended summer (June, July, August and September), as do 26% of the respiratory mortality.

Table 2. Mean number of deaths occurring during in months affected by large fires (LF) from June to September in the period 2001-2016 in the months of June, July, August and September 2001-2016 (total of 64 target summer months) in the 23 NUTS III sub-regions of mainland Portugal.

| Natural Deaths (N) | | Cardiovasc. deaths (N) | | Respiratory deaths (N) | | | | Inhabitants (2016) |
|--------------------|---------------------------|------------------------|--------------|------------------------|--------------|------------|--------------|--------------------|
| | NUTS III | All Months | Months w/ LF | All Months | Months w/ LF | All Months | Months w/ LF | |
| North | Alto Minho | 44434 | 13213 | 16586 | 4731 | 4975 | 1293 | 233813244149 |
| | Cávado | 44307 | 12832 | 14136 | 3931 | 5964 | 1471 | 404664411028 |
| | Ave | 49068 | 14153 | 15736 | 4421 | 5754 | 1363 | 415617425661 |
| | Alto Tâmega | 20527 | 5324 | 6687 | 1864 | 2298 | 576 | 8794193615 |
| | Terras de Trás-os-Montes | 24664 | 7190 | 8083 | 2312 | 2623 | 633 | 109409-116713 |
| | A.M. Porto | 221105 | 64366 | 67239 | 18660 | 24914 | 6252 | 17190211758991 |
| | Tâmega e Sousa | 50971 | 14623 | 18038 | 4847 | 6504 | 1574 | 420854432946 |
| | Douro | 39670 | 11648 | 13162 | 3642 | 4511 | 1143 | 193202204121 |
| Centre | <u>Região de</u> Aveiro | 53380 | 15440 | 17705 | 4945 | 6664 | 1647 | 363752369287 |
| | Viseu Dão-Lafões | 48379 | 14170 | 17700 | 4918 | 6474 | 1679 | 256928266207 |
| | <u>Região de</u> Coimbra | 81397 | 23648 | 28146 | 7718 | 10898 | 2807 | 439507-456871 |
| | Beiras e Serra da Estrela | 53414 | 15699 | 17816 | 5036 | 6108 | 1558 | 218961233478 |
| | <u>Região de</u> Leiria | 45190 | 13259 | 14294 | 4015 | 5507 | 1426 | 287770293941 |
| | Médio Tejo | 49769 | 13908 | 16666 | 4595 | 5417 | 1444 | 236256245940 |
| | Beira Baixa | 22545 | 6624 | 8063 | 2306 | 2267 | 562 | 8273188134 |
| | Oeste | 60896 | 17819 | 22467 | 6285 | 6539 | 1660 | 358029362311 |
| A.M. Lisboa | | 399704 | 118206 | 147172 | 41599 | 38931 | 10117 | 28213492827050 |
| Alentejo | Lezíria do Tejo | 45762 | 13505 | 16241 | 4491 | 5025 | 1397 | 239977247857 |
| | Alto Alentejo | 29145 | 8622 | 10391 | 2957 | 3690 | 923 | 108588117357 |
| | Alentejo Central | 33134 | 9602 | 12171 | 3252 | 2979 | 765 | 156207165688 |
| | Alentejo Litoral | 19241 | 5681 | 6916 | 1972 | 2294 | 635 | 9429197878 |
| | Baixo Alentejo | 30823 | 8988 | 11955 | 3317 | 3190 | 875 | 119024125875 |
| Algarve | | 71445 | 21715 | 22591 | 6482 | 7926 | 2307 | 441469446140 |

Considering that this period corresponds to one third of the year, the mortality in these months is slightly below what would be a regularly-distributed percentage for a 12-month timeframe (33%). This is in line with the evidence that cold temperatures are in general more responsible for all-cause deaths than warm ones (Lee et al., 2018; Scovronick et al., 2018).

305 Algarve, Alto Minho, Alto Alentejo and A.M. Lisbon are the NUTS with the higher percentage of all-cause mortality for the studied months, but the NUTS with more per capita incidence are Beira Baixa, Alto Alentejo, Baixo Alentejo and Beiras e Serra da Estrela, areas with lower population density and with mean higher age than the rest of the country.

310 Regarding With respect to cardiovascular mortality, the NUTS III which present a high incidence are Algarve, Terras de Trás-os-Montes and Beira Baixa, with the latter, Baixo Alentejo and Alto Alentejo having a higher percentage of population affected.

Finally, the results obtained for respiratory mortality show that Algarve, Lezíria-Lezíria do Tejo and Alentejo Litoral are the NUTS III which top the ranking in the summer months, whereas Alto Alentejo is the region with most population affected. Alentejo and Algarve suffer from high temperatures in the summer, which may also be an indicator that contribute for a higher mortality in general (Basu and Samet, 2002) (Basu and Samet, 2002) but also due to cardiovascular and respiratory diseases (Pinheiro et al., 2014). As is the (Pinheiro et al., 2014). In addition, the aforementioned regions suffer from a considerable afflux of tourists that increase their population in the same period, particularly in the Algarve. In Alentejo, the combination of high temperatures with an aged population and less health care resources available may be the justification to have the most population affected by mortality (Chen et al., 2019) (Chen et al., 2019). In fact, this is a tendency that has been becoming stronger since the beginning of the XXI century, as the percentage of population over 65 years-old changed in Alentejo from 315 22.5% in 2001 to 25.4% in 2018, higher than the percentages in the whole of Portugal (from 16.4% in 2001 to 21.7% in 2018) (as derived from PORDATA, <https://www.pordata.pt>).

The negative impact particulate matter can bring to human health is well established and can be translated into several types of diseases (Kim et al., 2015).

3.2.2 Associations between mortality and PM10

325 Wildfires are an important source of this pollutant particulate matter and the associations between wildfires and mortalitymortality, PM10, and the occurrence of large wildfires were assessed in this work.

330 section. As shown in Figure 6,a, four NUTS (three NUTS III (Alto Tâmega, Beiras e Serra da Estrela, and Viseu Dão-Lafões, Coimbra and Lezíria do Tejo)) present associations between wildfire-generated PM10 and all-cause mortality during the studied period, being these related to large fires in two of the NUTS, namely Lezíria do Tejo and Região de Coimbra. The None showed a direct significant association with the occurrence of large fires, likely due to the fact that their contribution to the total burned area in each year from 2001 to 2016 (Table 1) was highly variable (from 12.1% in 2011 to 79.5% in 2003). However, the wildfire origin of PM10 is corroborated by the positive significant correlations obtained for these three NUTS between PM10 and burnt area obtained for these four NUTS (Figure 5).

335 It is an evidence that particulate matter can enter the human body, arrive to the bloodstream and damage some organs or even provoke death due to cardiovascular afflictions like stroke or heart attack, among others, representing a clear hazard to public health (Brook et al., 2010; Hamanka and Mutlu, 2018). Regarding cardiovascular mortality, five NUTS present associations with PM10: Alto Minho, Douro, Região de Aveiro, Lezíria do Tejo and with Viseu Dão-Lafões displaying the highest correlations.

340 Beiras e Serra da Estrela, with only the last one being associated with the occurrence of large fires (Figure 6,b). This NUTS region was the is the NUTS region most affected by these events large wildfires during the studied period, both in number (\approx 50) and respective burned area ($>100,000$ ha).

345 Beiras e Serra da Estrela counts with almost 50 large fires during 2001–2016, which, and Viseu Dão-Lafões is the third in occurrences (28) corresponding to over 58,000 ha burned. This involves high levels of PM10 in a short period of time, which might provoke damage in human health, particularly in an aged population (e.g., for Beiras e Serra da Estrela, 23.8 and 28.7% over 65 years-old in 2001 and 2018, respectively; PORDATA, <https://www.pordata.pt>)).

350 Particulate matter also can damage the human respiratory system. The risk depends on the size of the particle, which if very small can even reach the alveolus (Neuberger et al., 2004; Jo et al., 2017). For respiratory mortality, four NUTS present In terms of types of diseases, for cardiovascular mortality five NUTS presented associations with PM10: Área Metropolitana do Porto, Douro, Alto Minho, A.M. Porto, Região de Aveiro and Viseu Dão-Lafões, all located in the north of the country, Região de Coimbra, and Algarve (Figure 6,e). The majority had a correlation with the presence b). Again, no direct significant associations were obtained with the occurrence of large fires, except Viseu Dão-Lafões.

355 Aveiro and Oporto regions are urbanized areas with considerable industrial presence and a cooler and rainier climate year-round than in the south, and the chronic exposure to these conditions suggests that people may be more susceptible to have their respiratory tract affected by acute PM-releasing events like large fires.

360 365 In some NUTS III regions like Tâmega e Sousa and Área Metropolitana de Lisboa no correlations were expected since there was not a high number of large fires. In addition, these fires did not spread through a very large area. On the other hand, for Alentejo NUTS (Alto Alentejo, Alentejo Central, Alentejo Litoral and Baixo Alentejo), Oeste, Leiria, Beira Baixa, Médio Tejo, Cávado, Ave, Terras de Trás os Montes and Alto Tâmega pollutant data were missing or there was only one station for the whole region. From these five NUTS, only Região de Aveiro showed a significant correlation between PM10 and burned area, revealing the impact of wildfires in the origin of the PM10. For respiratory mortality, only Viseu Dão-Lafões is one of the NUTS with less present associations with PM10 concentration for the whole Portugal mainland. Finally, for Alto Minho and Algarve no correlation was found between wildfires and, for which a strong correlation between PM10 and burned area was found, suggesting again the impact of wildfires on the presence of PM10 (Figure 6.c).

365 Portugal is a country that suffers constantly from serious wildfire incidents, which are bound to pose a risk not only to chronically affected populations but also from acute impacts of the pollutants released in such events. In this work, analysing the summer months (June to September) on a lengthy timeframe (2001–2016), it was possible to find significant associations between burned area and mortality in some NUTS III regions of mainland Portugal (mainly inland and in the north), as well as a significant correlation between burned area and PM10.

370 In particular, large fires (in this study considered above 1000 ha of burned area) have an impact on the health of the population in some areas due to the emission of particulate matter. Moreover, in such severe events, the population exposed to a high concentration of pollutants in a short period of time should be considered as a risk modifier of the impacts of air pollution exposure (Desikan, 2017; Rappold et al., 2017).

Relative risks (RR, numbers) obtained from Poisson regression for PM10 and (a) all-cause mortality; (b) cardiovascular mortality and (c) respiratory mortality from June to September in the period 2001–2016 (light blue NUTS III regions indicate a significant result, $p < 0.05$; dark blue represents significant result, $p < 0.05$, with fires >1000 ha; yellow NUTS III regions present no significant results for the variables studied; grey NUTS III regions indicate that no data was available for correlations). Only significant RR are displayed.

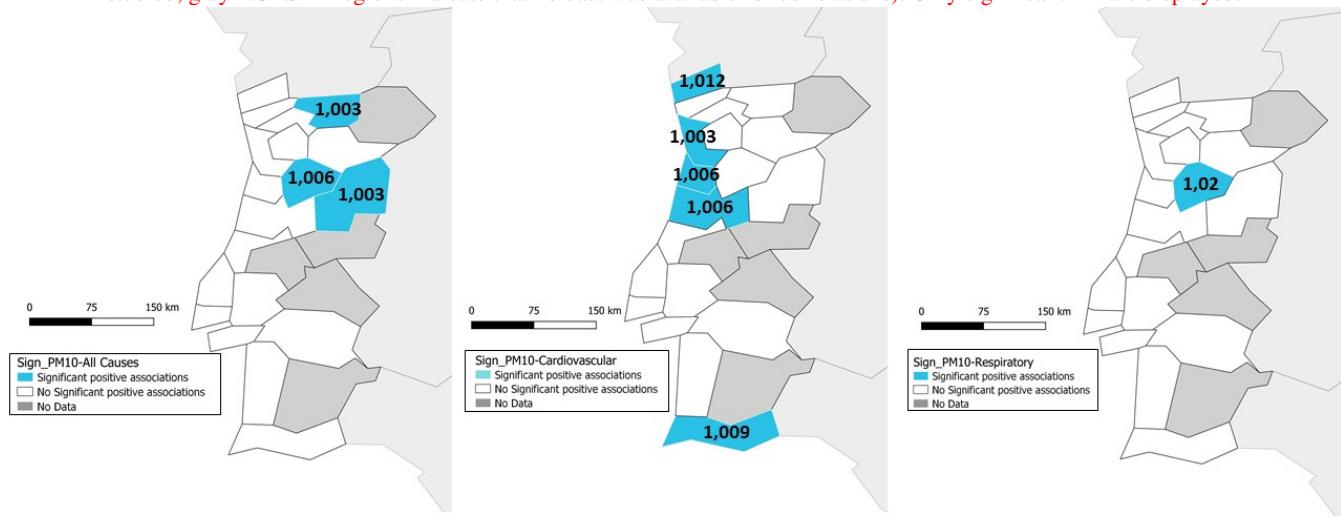


Figure 6. Relative risks (RR, numbers) obtained from Poisson regression for PM10 and (a) all-cause mortality; (b) cardiovascular mortality and (c) respiratory mortality from June to September in the period of 2001-2016 (light blue NUTS III regions indicate a significant result, $p < 0.05$; white NUTS III regions present no significant results for the variables studied; grey NUTS III regions indicate that no data was available for correlations). Only significant RR are shown.

4 Discussion

During the studied period (2001-2016), the 48% of wildfires occurred in Portugal were large fires, large fires were responsible for 46% of the more than 2 million ha of forest were burned in mainland Portugal. The areas that are more most affected by number and size of wildfires are north, centre center and inland of Portugal the country. Wildfires do not follow a pattern in number of the occurrences or size during the years studied. These evidences were This evidence was found despite the difficulties that the uneven scattering of the air quality monitoring stations analysing PM10 in Portugal posed, since. In fact, the areas where wildfires are usually more frequent (inland) are far from the urban centres (mainly along the coast), and thus, not abundant in air quality data availability due to the shortage (or even lack in some NUTS III) of monitoring stations. These regions also have an aged population, poorer economy and less health care resources, which can lead to an increase in the mortality rates in general. The socio-economic status of the population affected and the health care facilities and measures existing in the communities have to be taken into account (Oliveira et al., 2017), adding to the countless parameters that may affect these estimations which contribute to considerable gaps identified in this type of studies (Black et al., 2017).

385 Unfortunately, the scarce data available and the lack of accuracy in the existing ones prevented us from estimating/including a correction regarding their influence.

390 During the summer months occur almost 30% of the deaths of all diseases. Nevertheless, it was possible to find relationships between very relevant parameters. The significant positive correlation between PM10 and burned area found for 7 of the 13 NUTS III with available data is a good indication of where the influence of wildfires on the emission of PM10 is likely to be stronger. Although the location of the air quality monitoring stations may influence these correlations, especially when they are scarcer, for these NUTS it reveals the influence of wildfires on the local levels of PM10.

395 Large wildfires tend to be active for several days, releasing high amounts of pollutants to the atmosphere. In Portugal, such as in other Mediterranean countries, most of the wildfires are potentiated by strong winds, which may spread the fire smoke over large distances (Turco et al., 2019; Augusto et al., 2020). Thus, air quality monitoring stations located far from the ignition sites can still detect increases in the concentrations of, for instance, PM10. Augusto et al. (2020), while studying the impact of the uncontrolled wildfires of October 2017 in Portugal on the mortality, found that the PM10 and cardiovascular while for respiratory diseases mortality is around a 26%, so we can conclude that cold temperatures are in more responsible for all-cause deaths than warm ones. PM2.5 emitted reached the United Kingdom, as well as other northern European countries. Likewise, in Finland, where ambient PM levels are relatively low compared to other countries in Europe, 400 most of the strongest PM pollution episodes are typically related to emissions from wildfires in eastern European countries (Russia, Belarus, Ukraine, Estonia, Latvia and Lithuania) at a distance of hundreds to thousands of kilometres from southern Finland (Niemi et al., 2009; Kollanus et al., 2016).

405 The negative impact particulate matter can bring to human health is well established and can be translated into several types of diseases (Kim et al., 2015). It is an evidence that it can enter the human body, arrive to the bloodstream and damage some organs or even provoke death due to cardiovascular afflictions like stroke or heart attack, among others, representing a clear hazard to public health (Brook et al., 2010; Hamanaka and Mutlu, 2018). Particulate matter also can damage the human respiratory system. The risk depends on the size of the particle, which if very small can even reach the alveolus (Neuberger et al., 2004; Jo et

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410 In our study, the NUTS III where PM10 concentrations were found to be correlated with the burned area from large fires (Cávado, Ave, Tâmega e Sousa, Região de Aveiro, Viseu Dão-Lafões, Alto Tâmega and Beiras e Serra da Estrela) are indeed the ones where it would be expected to find the strongest influence of the wildfire originated PM10 on the population mortality. And although not for all, indeed associations between all-cause mortality and PM10 were found for three of these NUTS (Alto Tâmega, Viseu Dão-Lafões and Beiras e Serra da Estrela), with RR varying from 1.003 to 1.006; between cardiovascular mortality and PM10 for one (Região de Aveiro), with a RR of 1.006; and between respiratory mortality and PM10 also for one (Viseu Dão-Lafões) with a RR of 1.020. Associations between cardiovascular mortality and PM10 were found for four NUTS (Alto Minho, A.M. Porto, Coimbra and Algarve) where there was no significant correlation of PM10 with burned area. All these are located in the coast, where the population density in Portugal is clearly predominant (particularly in A.M. Porto), as well as considerable industrial presence. In these regions, cardiovascular disease may have many other sources, some of them derived from a more sedentary and stressed lifestyle.

420 The mortality increase associated with PM10 is consistent with the estimates reported in other European studies, such as APHEA2 (Katsouyanni et al., 2001), APHENA (Samoli et al., 2008), EpiAir (Faustini et al., 2011) and MED-PARTICLES (Faustini et al., 2015), which also reported higher PM10 effects on all-cause, cardiovascular, and respiratory mortalities.

425 However, some studies present uneven conclusions. Johnston et al. (2011) reported the highest effects on cardiovascular mortality, but Morgan et al. (2010) did not find any consistent effect with cardiovascular deaths in Australia, and Analitis et al. (2012) registered the highest effects on respiratory mortality in Greece. This high variability may be related to several factors, notably: i) different PM composition or varying gaseous emissions (CO, VOCs, NOx or SO₂) from wildfires, which may have different degrees of toxicity on cardiovascular and respiratory systems; or ii) increasing temperature during wildfires, which is known to enhance the effects of PM on more susceptible individuals (e.g., cardiac patients) (Qian et al., 2008). Therefore, the effects we found on all-cause, cardiovascular, and respiratory mortalities during the wildfire seasons may be due to different PM 430 compositions or increasing temperature.

435 Region-specific associations between PM10 concentrations and mortality were also observed. These may have been influenced by the factors described above (different PM composition and increasing temperature), but also by the magnitude and duration of the exposure to PM from a given fire; the underlying health status of the population; and the size of the population. The age of the exposed individuals can also be important. In some studies, larger effect estimates in groups of 65 years and older 440 have been reported. Analitis et al. (2012) mentioned that the effect of respiratory mortality in Greece was higher in adults of ages 75 and above during large fires, whereas Haikerwal et al. (2015) observed an increase in risk of cardiac arrests, especially in older adults in Australia, although not all resulted in death. In Brazil, Nunes et al. (2013) reported that older adults had the strongest association between exposure to biomass burning and circulatory disease mortality. In Portugal, the regions traditionally impacted by wildfires coincide with a larger percentage of aged population, which can help explaining the obtained associations.

445 In our analysis, to study the relationship between the burned area and PM10 concentrations, averaged monthly data were used, as the minimum temporal scale available for the burned area was one month. Other studies relating wildfire-originated PM and mortality are usually based on daily PM concentrations and daily death counts, since they do not account for the burned area as a measure of the wildfire size. Therefore, the monthly approach obviously reduced the number of data available 450 and the possibility of finding more significant correlations, which may have diluted the effects of some wildfires on the population mortality. Moreover, some health effects may not have been detected because wildfires are episodic and local events. Nevertheless, the results provide an overall context, highlighting the strongest associations between wildfire generated PM10 and all-cause, cardiovascular, and respiratory mortalities. Being able to achieve them with this uneven distribution of available data is an indication that the approach can be very useful to at least uncover tendencies and, in regions with stronger monitoring capabilities and coverage, a way to find stronger and more accurate correlations. This will help legislators and other government bodies to propose ways to protect the population chronically exposed to wildfires or more susceptible to acute reactions to wildfire smoke.

5 Conclusions

Portugal is a country that suffers constantly from serious wildfire incidents, which are bound to pose a risk not only to 455 chronically affected populations but also from acute impacts of the pollutants released in such events. In this work, analysing the summer months (June to September) on a lengthy timeframe (2001-2016), it was possible to find relevant associations between PM10 (associated with large wildfires) and mortality in some NUTS III regions of mainland Portugal (mainly inland and in the north), as well as a significant correlation between burned area and PM10.

In particular, it was found that large fires (in this study considered above 1,000 ha of burned area) have an impact on the 460 health of the population in some areas due to the emission of particulate matter. The lack of data or possible confounding factors likely prevented a higher number of NUTS III with significant correlations. Moreover, in such severe events, the population exposed to a high concentration of pollutants in a short period of time should be considered as a risk modifier of the impacts of air pollution exposure (Desikan, 2017; Rappold et al., 2017).

These episodes occurred during the summer months (June-July-August-September), when high temperatures and long 465 episodes of drought increase the probabilities of undergoing one of these extreme events. On a future ruled by climate changes, the high temperatures and long periods of drought that usually fuel big fires are expected to increase, thus leading the way for more extreme and intense events to occur, even outside the typically affected regions. Thus, more population will be exposed more frequently to high pollutant levels, affecting their general health, and increasing chronic diseases and mortality. Hence, 470 restrictive policies and protocols to improve the effectiveness of preventive and mitigation actions must be enforced to face this environmental and societal issue.

Data availability. Data is publicly available through the websites mentioned in the text:

- EFFIS (European Forest Fire Information System). Data and Services, 2019. Available at <https://effis.jrc.ec.europa.eu/applications/data-and-services/> (last accessed 11/10/2019).
- ICNF (Instituto da Conservação da Natureza e das Florestas), 2019. Available at <https://www.icnf.pt/> (last accessed: 01/09/2019).
- 475 INE (Instituto Nacional de Estatística). Statistics Portugal - Web Portal, 2019. Available at: https://www.ine.pt/xportal/xmain?xpid=INE&xpgid=ine_indicadores&contexto=pi&indOcorrCod=0008273&selTab=tab0 (last accessed 16/10/2019).
- GWF (Global Forest Watch), 2020. Available at <https://www.globalforestwatch.org/map/?gfwfires=true> (last accessed, 10/10/2020).
- PORDATA (Base de Dados Portugal Contemporâneo). População residente: total e por grandes grupos etários (in Portuguese), 2019. Available at <https://www.pordata.pt> (last accessed 31/03/2020).

480 All the compiled data is available upon contacting the corresponding author (pedro.jimenezguerrero@um.es)

Author contributions. PT-C wrote the manuscript, with contributions from SA and NR. The manuscript was finally revised by PJ-G. P-TC and SA designed the experiments and led the statistical analysis, with the support of LP-P, NR and PJ-G.

Competing interests. The authors declare no conflict of interest.

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References

490 Analitis, A., Georgiadis, I., and Katsouyanni, K.: Forest fires are associated with elevated mortality in a dense urban setting, *Occupational and Environmental Medicine*, 69, 158–162, <https://doi.org/10.1136/oem.2010.064238>, 2012.

Anderson, J. O., Thundiyil, J. G., and Stolbach, A.: Clearing the Air: A Review of the Effects of Particulate Matter Air Pollution on Human Health, *Journal of Medical Toxicology*, 8, 166–175, <https://doi.org/10.1007/s13181-011-0203-1>, 2012.

495 Augusto, S., Ratola, N., Tarín-Carrasco, P., Jiménez-Guerrero, P., Turco, M., Schuhmacher, M., Costa, S., Teixeira, J., and Costa, C.: Population exposure to particulate-matter and related mortality due to the Portuguese wildfires in October 2017 driven by storm Ophelia, *Environment International*, 144, 106 056, <https://doi.org/10.1016/j.envint.2020.106056>, 2020.

Basu, R. and Samet, J. M.: Relation between Elevated Ambient Temperature and Mortality: A Review of the Epidemiologic Evidence, *Epidemiologic Reviews*, 24, 190–202, <https://doi.org/10.1093/epirev/mxf007>, 2002.

500 Black, C., Tesfaigzi, Y., Bassein, J. A., and Miller, L. A.: Wildfire smoke exposure and human health: Significant gaps in research for a growing public health issue, *Environmental Toxicology and Pharmacology*, 55, 186 – 195, <https://doi.org/https://doi.org/10.1016/j.etap.2017.08.022>, 2017.

Bowman, D. M. J. S., Williamson, G. J., Abatzoglou, J. T., Kolden, C. A., Cochrane, M. A., and Smith, A. M. S.: Human exposure and sensitivity to globally extreme wildfire events, *Nature Ecology & Evolution*, 1, 58, <https://doi.org/10.1038/s41559-016-0058>, 2017.

505 Brook, R. D., Rajagopalan, S., Pope, C. A., Brook, J. R., Bhatnagar, A., Diez-Roux, A. V., Holguin, F., Hong, Y., Luepker, R. V., Mittleman, M. A., Peters, A., Siscovich, D., Smith, S. C., Whitsel, L., and Kaufman, J. D.: Particulate Matter Air Pollution and Cardiovascular Disease: an update to the scientific statement from the American Heart Association, *Circulation*, 121, 2331–2378, <https://doi.org/10.1161/CIR.0b013e3181dbece1>, 2010.

Cascio, W. E.: Wildland fire smoke and human health, *Science of The Total Environment*, 624, 586 – 595, <https://doi.org/https://doi.org/10.1016/j.scitotenv.2017.12.086>, 2018.

510 Chen, C.-H., Wu, C.-D., Chiang, H.-C., Chu, D., Lee, K.-Y., Lin, W.-Y., Yeh, J.-I., Tsai, K.-W., and Guo, Y.-L. L.: The effects of fine and coarse particulate matter on lung function among the elderly, *Scientific Reports*, 9, 14 790, <https://doi.org/10.1038/s41598-019-51307-5>, 2019.

Ciscar, J.-C., Feyen, L., Soria, A., Lavalle, C., Raes, F., Perry, M., Nemry, F., Demirel, H., Rozsai, M., Dosio, A., et al.: Climate impacts in Europe-The JRC PESETA II project, in: *JRC Scientific and Policy Reports EUR 26586EN*, vol. 26586, p. 155, Joint Research Center, 515 2014.

Desikan, A.: Outdoor air pollution as a possible modifiable risk factor to reduce mortality in post-stroke population, *Neural Regeneration Research*, 12, 351–353, <https://doi.org/10.4103/1673-5374.202917>, 2017.

Eurostat: NUTS 2016, version 14/03/2019, 1:60 million scale, Retrieved at <https://ec.europa.eu/eurostat/web/gisco/geodata/reference-data/administrative-units-statistical-units/nuts> (last accessed on 19 June 2021), 2019.

520 Faustini, A., Stafoggia, M., Berti, G., Bisanti, L., Chiusolo, M., Cernigliaro, A., Mallone, S., Primerano, R., Scarnato, C., Simonato, L., Vigotti, M., and Forastiere, F.: The relationship between ambient particulate matter and respiratory mortality: a multi-city study in Italy, *European Respiratory Journal*, 38, 538–547, <https://doi.org/10.1183/09031936.00093710>, 2011.

Faustini, A., Alessandrini, E. R., Pey, J., Perez, N., Samoli, E., Querol, X., Cadum, E., Perrino, C., Ostro, B., Ranzi, A., Sunyer, J., Stafoggia, M., and Forastiere, F.: Short-term effects of particulate matter on mortality during forest fires in Southern Europe: results of the MED-525 PARTICLES Project, *Occupational and Environmental Medicine*, 72, 323–329, <https://doi.org/10.1136/oemed-2014-102459>, 2015.

Fuller, G. W., Tremper, A. H., Baker, T. D., Yttri, K. E., and Butterfield, D.: Contribution of wood burning to PM10 in London, *Atmospheric Environment*, 87, 87 – 94, <https://doi.org/10.1016/j.atmosenv.2013.12.037>, 2014.

Gillett, N. P., Weaver, A. J., Zwiers, F. W., and Flannigan, M. D.: Detecting the effect of climate change on Canadian forest fires, *Geophysical Research Letters*, 31, L18 211, <https://doi.org/10.1029/2004GL020876>, 2004.

530 Haikerwal, A., Akram, M., Monaco, A. D., Smith, K., Sim, M. R., Meyer, M., Tonkin, A. M., Abramson, M. J., and Dennekamp, M.: Impact of Fine Particulate Matter (PM2.5) Exposure During Wildfires on Cardiovascular Health Outcomes, *Journal of the American Heart Association*, 4, e001 653, <https://doi.org/10.1161/JAHA.114.001653>, 2015.

Hamanaka, R. B. and Mutlu, G. M.: Particulate Matter Air Pollution: Effects on the Cardiovascular System, *Frontiers in Endocrinology*, 9, 680, <https://doi.org/10.3389/fendo.2018.00680>, 2018.

535 Hoek, G., Krishnan, R. M., Beelen, R., Peters, A., Ostro, B., Brunekreef, B., and Kaufman, J. D.: Long-term air pollution exposure and cardio-respiratory mortality: A review, *Environmental Health: A Global Access Science Source*, 12, <https://doi.org/10.1186/1476-069X-12-43>, 2013.

540 Im, U., Brandt, J., Geels, C., Hansen, K. M., Christensen, J. H., Andersen, M. S., Solazzo, E., Kioutsoukis, I., Alyuz, U., Balzarini, A., Baro, R., Bellasio, R., Bianconi, R., Bieser, J., Colette, A., Curci, G., Farrow, A., Flemming, J., Fraser, A., Jimenez-Guerrero, P., Kitwiroon, N., Liang, C.-K., Nopmongcol, U., Pirovano, G., Pozzoli, L., Prank, M., Rose, R., Sokhi, R., Tuccella, P., Unal, A., Vivanco, M. G., West, J., Yarwood, G., Hogrefe, C., and Galmarini, S.: Assessment and economic valuation of air pollution impacts on human health over Europe and the United States as calculated by a multi-model ensemble in the framework of AQMEII3, *Atmospheric Chemistry and Physics*, 18, 5967–5989, <https://doi.org/10.5194/acp-18-5967-2018>, 2018.

Islam, M. A. and Chowdhury, R. I.: A generalized right truncated bivariate Poisson regression model with applications to health data, *PLOS One*, 12, 1–13, <https://doi.org/10.1371/journal.pone.0178153>, 2017.

545 Jo, E.-J., Lee, W.-S., Jo, H.-Y., Kim, C.-H., Eom, J.-S., Mok, J.-H., Kim, M.-H., Lee, K., Kim, K.-U., Lee, M.-K., and Park, H.-K.: Effects of particulate matter on respiratory disease and the impact of meteorological factors in Busan, Korea, *Respiratory Medicine*, 124, 79–87, <https://doi.org/10.1016/j.rmed.2017.02.010>, 2017.

550 Johnston, F., Hanigan, I., Henderson, S., Morgan, G., and Bowman, D.: Extreme air pollution events from bushfires and dust storms and their association with mortality in Sydney, Australia 1994–2007, *Environmental Research*, 111, 811–816, <https://doi.org/10.1016/j.envres.2011.05.007>, 2011.

Johnston, F. H., Henderson, S. B., Chen, Y., Randerson, J. T., Marlier, M., DeFries, R. S., Kinney, P., Bowman, D. M., and Brauer, M.: Estimated Global Mortality Attributable to Smoke from Landscape Fires, *Environmental Health Perspectives*, 120, 695–701, <https://doi.org/10.1289/ehp.1104422>, 2012.

555 Katsouyanni, K., Touloumi, G., Samoli, E., Gryparis, A., Tertre, A. L., Monopolis, Y., Rossi, G., Zmirou, D., Ballester, F., Boumghar, A., Anderson, H. R., Wojtyniak, B., Paldy, A., Braunstein, R., Pekkanen, J., Schindler, C., and Schwartz, J.: Confounding and Effect Modification in the Short-Term Effects of Ambient Particles on Total Mortality: Results from 29 European Cities within the APHEA2 Project, *Epidemiology*, 12, 521–531, 2001.

Kim, K.-H., Kabir, E., and Kabir, S.: A review on the human health impact of airborne particulate matter, *Environment International*, 74, 136–143, <https://doi.org/10.1016/j.envint.2014.10.005>, 2015.

560 Knorr, W., Dentener, F., Hantson, S., Jiang, L., Klimont, Z., and Arneth, A.: Air quality impacts of European wildfire emissions in a changing climate, *Atmospheric Chemistry and Physics*, 16, 5685–5703, <https://doi.org/10.5194/acp-16-5685-2016>, 2016.

Kollanus, V., Tiittanen, P., Niemi, J. V., and Lanki, T.: Effects of long-range transported air pollution from vegetation fires on daily mortality and hospital admissions in the Helsinki metropolitan area, Finland, *Environmental Research*, 151, 351–358, 565 <https://doi.org/https://doi.org/10.1016/j.envres.2016.08.003>, 2016.

Lazaridis, M., Latos, M., Aleksandropoulou, V., Hov, Ø., Papayannis, A., and Tørseth, K.: Contribution of forest fire emissions to atmospheric pollution in Greece, *Air Quality, Atmosphere & Health*, 1, 143–158, <https://doi.org/10.1007/s11869-008-0020-0>, 2008.

Lee, W., Choi, H. M., Lee, J. Y., Kim, D. H., Honda, Y., and Kim, H.: Temporal changes in mortality impacts of heat wave and cold spell in Korea and Japan, *Environment International*, 116, 136 – 146, <https://doi.org/https://doi.org/10.1016/j.envint.2018.04.017>, 2018.

Liang, C.-K., West, J. J., Silva, R. A., Bian, H., Chin, M., Davila, Y., Dentener, F. J., Emmons, L., Flemming, J., Folberth, G., Henze, D., Im, U., Jonson, J. E., Keating, T. J., Kucsera, T., Lenzen, A., Lin, M., Lund, M. T., Pan, X., Park, R. J., Pierce, R. B., Sekiya, T., Sudo, K., and Takemura, T.: HTAP2 multi-model estimates of premature human mortality due to intercontinental transport of air pollution and emission sectors, *Atmospheric Chemistry and Physics*, 18, 10 497–10 520, <https://doi.org/10.5194/acp-18-10497-2018>, 2018.

Lidskog, R., Johansson, J., and Sjödin, D.: Wildfires, responsibility and trust: public understanding of Sweden's largest wildfire, *Scandinavian Journal of Forest Research*, 34, 319–328, <https://doi.org/10.1080/02827581.2019.1598483>, 2019.

Lin, M., Fiore, A. M., Horowitz, L. W., Cooper, O. R., Naik, V., Holloway, J., Johnson, B. J., Middlebrook, A. M., Oltmans, S. J., Pollack, I. B., Ryerson, T. B., Warner, J. X., Wiedinmyer, C., Wilson, J., and Wyman, B.: Transport of Asian ozone pollution into surface air over the western United States in spring, *Journal of Geophysical Research: Atmospheres*, 117, D00V07, <https://doi.org/https://doi.org/10.1029/2011JD016961>, 2012.

Liu, J. C., Pereira, G., Uhl, S. A., Bravo, M. A., and Bell, M. L.: A systematic review of the physical health impacts from non-occupational exposure to wildfire smoke, *Environmental Research*, 136, 120 – 132, <https://doi.org/https://doi.org/10.1016/j.envres.2014.10.015>, 2015.

Maia, P., Keizer, J., Vasques, A., Abrantes, N., Roxo, L., Fernandes, P., Ferreira, A., and Moreira, F.: Post-fire plant diversity and abundance in pine and eucalypt stands in Portugal: Effects of biogeography, topography, forest type and post-fire management, *Forest Ecology and Management*, 334, 154 – 162, <https://doi.org/https://doi.org/10.1016/j.foreco.2014.08.030>, 2014.

Moreira, F., Viedma, O., Arianoutsou, M., Curt, T., Koutsias, N., Rigolot, E., Barbati, A., Corona, P., Vaz, P., Xanthopoulos, G., Mouillet, F., and Bilgili, E.: Landscape – wildfire interactions in southern Europe: Implications for landscape management, *Journal of Environmental Management*, 92, 2389 – 2402, <https://doi.org/https://doi.org/10.1016/j.jenvman.2011.06.028>, 2011.

Morgan, G., Sheppeard, V., Khalaj, B., Ayyar, A., Lincoln, D., Jalaludin, B., Beard, J., Corbett, S., and Lumley, T.: Effects of Bushfire Smoke on Daily Mortality and Hospital Admissions in Sydney, Australia, *Epidemiology*, 21, 2010.

Münzel, T., Gori, T., Al-Kindi, S., Deanfield, J., Lelieveld, J., Daiber, A., and Rajagopalan, S.: Effects of gaseous and solid constituents of air pollution on endothelial function, *European Heart Journal*, 39, 3543–3550, <https://doi.org/10.1093/eurheartj/ehy481>, 2018.

Neuberger, M., Schimek, M. G., Horak, F., Moshammer, H., Kundi, M., Frischer, T., Gomiscek, B., Puxbaum, H., and Hauck, H.: Acute effects of particulate matter on respiratory diseases, symptoms and functions: epidemiological results of the Austrian Project on Health Effects of Particulate Matter (AUPHEP), *Atmospheric Environment*, 38, 3971–3981, 595 <https://doi.org/https://doi.org/10.1016/j.atmosenv.2003.12.044>, 2004.

Niemi, J. V., Saarikoski, S., Aurela, M., Tervahattu, H., Hillamo, R., Westphal, D. L., Aarnio, P., Koskentalo, T., Makkonen, U., Vehkamäki, H., and Kulmala, M.: Long-range transport episodes of fine particles in southern Finland during 1999–2007, *Atmospheric Environment*, 43, 1255–1264, <https://doi.org/https://doi.org/10.1016/j.atmosenv.2008.11.022>, 2009.

Nunes, K. V. R., Ignotti, E., and Hacon, S. d. S.: Circulatory disease mortality rates in the elderly and exposure to PM2.5 generated by biomass
600 burning in the Brazilian Amazon in 2005, *Cadernos de Saúde Pública*, 29, 589–598, <https://doi.org/10.1590/S0102-311X2013000300016>,
2013.

Nunes, L. J., Meireles, C. I., Gomes, C. J., and Ribeiro, N. M.: Historical development of the Portuguese forest: The introduction of invasive
605 species, *Forests*, 10, <https://doi.org/10.3390/f10110974>, 2019.

Oliveira, S., Zêzere, J. L., Queirós, M., and Pereira, J. M.: Assessing the social context of wildfire-affected areas. The case of mainland
610 Portugal, *Applied Geography*, 88, 104 – 117, <https://doi.org/https://doi.org/10.1016/j.apgeog.2017.09.004>, 2017.

Pallarés, S., Gómez, E., Martínez, A., and Jordán, M. M.: The relationship between indoor and outdoor levels of PM10 and its chemical
615 composition at schools in a coastal region in Spain, *Heliyon*, 5, e02270, <https://doi.org/https://doi.org/10.1016/j.heliyon.2019.e02270>, 2019.

Parente, J., Pereira, M., Amraoui, M., and Tedim, F.: Negligent and intentional fires in Portugal: Spatial distribution characterization, *Science
620 of The Total Environment*, 624, 424 – 437, <https://doi.org/https://doi.org/10.1016/j.scitotenv.2017.12.013>, 2018.

Pearson, K. and Galton, F.: VII. Note on regression and inheritance in the case of two parents, *Proceedings of the Royal Society of London*,
58, 240–242, <https://doi.org/10.1098/rspl.1895.0041>, 1895.

Perrino, C., Tofful, L., Torre, S. D., Sargolini, T., and Canepari, S.: Biomass burning contribution to PM10 concentration in Rome (Italy): Sea-
625 sonal, daily and two-hourly variations, *Chemosphere*, 222, 839 – 848, <https://doi.org/https://doi.org/10.1016/j.chemosphere.2019.02.019>,
2019.

Pinheiro, S. d. L. L. d. A., Saldiva, P. H. N., Schwartz, J., and Zanobetti, A.: Isolated and synergistic effects of PM10 and average temperature
630 on cardiovascular and respiratory mortality, *Revista de Saúde Pública*, 48, 881–888, <https://doi.org/10.1590/S0034-8910.2014048005218>,
2014.

Qian, Z., He, Q., Lin, H.-M., Kong, L., Bentley, C. M., Liu, W., and Zhou, D.: High Temperatures Enhanced Acute Mortality Ef-
635 fects of Ambient Particle Pollution in the ‘Oven’ City of Wuhan, China, *Environmental Health Perspectives*, 116, 1172–1178,
<https://doi.org/10.1289/ehp.10847>, 2008.

Rahman, A., Luo, C., Khan, M. H. R., Ke, J., Thilakanayaka, V., and Kumar, S.: Influence of atmospheric PM2.5, PM10, O3, CO, NO2,
640 SO2, and meteorological factors on the concentration of airborne pollen in Guangzhou, China, *Atmospheric Environment*, 212, 290 – 304,
<https://doi.org/https://doi.org/10.1016/j.atmosenv.2019.05.049>, 2019.

Rappold, A. G., Cascio, W. E., Kilaru, V. J., Stone, S. L., Neas, L. M., Devlin, R. B., and Diaz-Sanchez, D.: Cardio-respiratory outcomes
645 associated with exposure to wildfire smoke are modified by measures of community health, *Environmental Health: A Global Access
Science Source*, 11, 1–9, <https://doi.org/10.1186/1476-069X-11-71>, 2012.

Rappold, A. G., Reyes, J., Pouliot, G., Cascio, W. E., and Diaz-Sanchez, D.: Community Vulnerability to Health Impacts of Wildland Fire
650 Smoke Exposure, *Environmental Science & Technology*, 51, 6674–6682, <https://doi.org/10.1021/acs.est.6b06200>, 2017.

Reid, C. E., Brauer, M., Johnston, F. H., Jerrett, M., Balmes, J. R., and Elliott, C. T.: Critical review of health impacts of wildfire smoke
655 exposure, *Environmental Health Perspectives*, 124, 1334–1343, <https://doi.org/10.1289/ehp.1409277>, 2016.

Rovira, J., Domingo, J. L., and Schuhmacher, M.: Air quality, health impacts and burden of disease due to air pollution (PM10, PM2.5,
660 NO2 and O3): Application of AirQ+ model to the Camp de Tarragona County (Catalonia, Spain), *Science of The Total Environment*, 703,
135 538, <https://doi.org/https://doi.org/10.1016/j.scitotenv.2019.135538>, 2020.

635 Samoli, E., Peng, R., Ramsay, T., Pipikou, M., Touloumi, G., Dominici, F., Burnett, R., Cohen, A., Krewski, D., Samet, J., and Katsouyanni, K.: Acute Effects of Ambient Particulate Matter on Mortality in Europe and North America: Results from the APHENA Study, Environmental Health Perspectives, 116, 1480–1486, <https://doi.org/10.1289/ehp.11345>, 2008.

640 San-Miguel-Ayanz, J., Durrant, T., Boca, R., Libertà, G., Branco, A., De Rigo, D., Ferrari, D., Maianti, P., Artés Vivancos, T., Costa, H., Schulte, E., and Loffler, P.: Forest Fires in Europe, Middle East and North Africa 2017, EUR 28707 EN, Publications Office of the European Union, Luxembourg, doi:10.2760/663443, 2017.

Schär, C., Vidale, P. L., Lüthi, D., Frei, C., Häberli, C., Liniger, M. A., and Appenzeller, C.: The role of increasing temperature variability in European summer heatwaves, *Nature*, 427, 332–336, <https://doi.org/10.1038/nature02300>, 2004.

645 Scovronick, N., Sera, F., Acquaotta, F., Garzena, D., Fratianni, S., Wright, C. Y., and Gasparrini, A.: The association between ambient temperature and mortality in South Africa: A time-series analysis, *Environmental Research*, 161, 229 – 235, <https://doi.org/https://doi.org/10.1016/j.envres.2017.11.001>, 2018.

Settele, J., Scholes, R., Betts, R. A., Bunn, S., Leadley, P., Nepstad, D., Overpeck, J., Taboada, M. A., Fischlin, A., Moreno, J. M., et al.: Terrestrial and inland water systems, in: *Climate change 2014 impacts, adaptation and vulnerability: Part A: Global and sectoral aspects*, pp. 271–360, Cambridge University Press, 2015.

650 Tarín-Carrasco, P., Morales-Suarez-Varela, M., Im, U., Brandt, J., Palacios-Peña, L., and Jiménez-Guerrero, P.: Isolating the climate change impacts on air-pollution-related-pathologies over central and southern Europe - A modelling approach on cases and costs, *Atmospheric Chemistry and Physics*, 19, 9385–9398, <https://doi.org/10.5194/acp-19-9385-2019>, 2019.

Trentmann, J., Yokelson, R. J., Hobbs, P. V., Winterrath, T., Christian, T. J., Andreae, M. O., and Mason, S. A.: An analysis of the chemical processes in the smoke plume from a savanna fire, *Journal of Geophysical Research: Atmospheres*, 110, D12301, <https://doi.org/https://doi.org/10.1029/2004JD005628>, 2005.

655 Turco, M., Jerez, S., Augusto, S., Tarín-Carrasco, P., Ratola, N., Jiménez-Guerrero, P., and Trigo, R. M.: Climate drivers of the 2017 devastating fires in Portugal, *Scientific Reports*, 9, 13 886, <https://doi.org/10.1038/s41598-019-50281-2>, 2019.

World Health Organization: Wildfires and heat-wave in the Russian Federation, in: *Public health advice*, p. 15, WHO Regional Office for Europe, Copenhagen, Denmark, 2010.