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# Land cover changes and forest landscape evolution (1985–2009) in a typical Mediterranean agroforestry system (High Agri Valley)

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## Abstract

The present study focuses on the transformations of a typical Mediterranean agro-forestry landscape of southern Italy (High Agri Valley – Basilicata region) occurred during 24 years. In this period, the valuable agricultural and natural areas that compose such a landscape were subjected to intensive industry-related activities linked to the exploitation of the largest European on-shore oil reservoir. Landsat imagery acquired in 1985 and 2009 were used to detect changes in forest areas and major land use trajectories. Landscape metrics indicators were adopted to characterize landscape structure and evolution of both the complex ecomosaic (14 land cover classes) and the Forest/Non Forest arrangement. Our results indicate a net increase of 11 % of forest areas between 1985 and 2009. The major changes concern: increase of all forest covers at the expense of pastures and grasses, enlargement of riparian vegetation, expansion of artificial areas. The observed expansion of forests was accompanied by a decrease of the fragmentation levels likely due to the reduction of small glades that break forest homogeneity and to the recolonization of herbaceous areas. Overall, we observe an evolution towards a more stable configuration depicting a satisfactory picture of vegetation health.

## 1 Introduction

The distinctive character of Mediterranean agro-forestry systems is the complex assemblage of different land covers resulting from a millenary history of man-made modifications (Antrop, 2004). Generally these are composed of agricultural fields, sparsely urbanized areas (roads, rural districts, etc.), and heterogeneous forests where a variety of vegetation types are aggregated according to diverse extents, shapes, and connectivity levels (Lafortezza et al., 2010). In addition, in the last decades the traditional composition and arrangement of Mediterranean landscapes have been significantly changed by urbanization, industrialization, logging, agricultural policies (e.g., European

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Common Agricultural Policy) and climate change effects (Grove and Rackham, 2001; Tanrivermis, 2003; Weber et al., 2005; FAO, 2014; Salvati et al., 2014).

In this context it is important to develop sustainable land management strategies that incorporate forest planning. Since the nineteenth century the main conceptual approach to forest planning considers forest landscapes not only in the perspective of providing wood and timber (Farcy, 2004) but as a precious container of several ecosystem services and functions (conservation of biodiversity, hydrogeological risk protection, centre of tourism and recreational activities, production of non-wood products such as mushrooms, fodder, honey, etc.) to be sustained over time. Therefore the definition of a framework of sound principles to plan a sustainable governance of forest resources holds a key role.

The new emerging paradigm, the so-called SFM-Sustainable Forest Management (Szaro et al., 2000), is based on the integration of multiple forest values. In Italy, Forest Landscape Management Plans (FLMP) are recently gaining importance as intermediate tools between national/regional forest management plans and forest regulations applied at farm level (Paletto et al., 2012).

Effective tools to support these plans have to provide updated observations to characterize forest evolution. Satellite data have a well-established history in informing forest management/inventory. In particular, Landsat imagery have been largely applied to forest monitoring and studies because they represent a trade off among the need of high spatial resolution (30 m), reasonable revisiting frequency (16 days), and availability of long time series (since July 1972) to carry out multitemporal analyses (Collins and Woodcock, 1996; Borrelli et al., 2014; Mancino et al., 2014).

The evolution of forests can be profitably explored adopting a landscape ecological approach (Laforteza et al., 2013) relied on the identification of ecological patterns generated by mutual interactions between anthropogenic activities and natural processes (Ingegnoli, 2002). By using landscape metrics it is possible to quantify parameters linked to geographical patterns generated by land-use/land cover dynamics. In fact, metrics are used in a wide range of disciplines with different applications, such as

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water quality (Uemaa et al., 2007), biodiversity (Schindler et al., 2008), land degradation (Imbrenda et al., 2014). Recently they have been successfully applied also to evaluate forest health status (State of Europe's Forests, FOREST EUROPE, UNECE and FAO 2011).

In particular, information obtained from landscape metrics can be very precious to improve silvicultural prescriptions and baselines for forest management, as well as to highlight effects of implemented actions (Sano et al., 2009; Reddy et al., 2013; Costa et al., 2014; Paliwal and Mathur; 2014).

The multitemporal analysis performed in this study (1985–2009) attempted to investigate changes occurred in the landscapes of the High Agri Valley (Basilicata – southern Italy), a typical example of Mediterranean landscape. This area is characterized by a consistent presence of agricultural activities in prevalently hilly environments and by a unique forest heritage with a rich biodiversity that have stimulated tourism growth. Contextually, a significant increase in anthropogenic disturbance was experienced since the Nineties due to the exploitation of the largest European on-shore oil reservoir and to the consequent induced activities.

Due to these peculiar conditions our purpose was to analyze the landscape structure and evolution of the area during the investigated period with a focus on forests.

To this aim land cover maps (14 classes) were created for the reference years and subsequently recoded in binary Forest/Non Forest maps. Then, in order to evaluate the ecological stability of the area we analyzed the results obtained by applying landscape metrics indicators on the above-mentioned maps.

Especially when used in multitemporal studies (Fichera et al., 2012), landscape metrics can capture the land use processes potentially connected to adverse impacts (forest loss and fragmentation), or positive environmental conditions (forest gain and increasing forest connectivity).

Additionally, we used metrics to evaluate the effects of different managements analyzing public and private ownership areas separately so as to understand how forest tenure regimes may influence arrangement, spatial heterogeneity and ecological

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functionality of vegetated classes (Botequilha Leitato and Ahern, 2002; Blondel et al., 2010; Laforteza et al., 2013).

The information obtained from these analyses gives us useful insights to support policy actions aiming to reconcile the increasing need of exploiting ecosystem services with the environmental conservation.

## 2 Materials

### 2.1 Study area

The high valley of the Agri river is an intermountain basin located in the southern Apennines (Basilicata, Italy), covering about 725 km<sup>2</sup> (Fig. 1).

The local climate is characterized by cold-humid winters and hot dry summers, as occurs in typical Mediterranean regions ([http://www1.inea.it/otris/file/clima\\_txt.htm](http://www1.inea.it/otris/file/clima_txt.htm), last access: 10 July 2014). According to Cantore et al. (1987) the total amount of annual rainfall is around 1100 mm, which would be sufficient on the average to allow for normal plant life. However precipitation shows a high annual excursion (more than 800 mm) and summer droughts are likely responsible for the physiognomy of vegetation types, sometimes predisposing them to parasitic attacks.

The area is characterized by human settlements that are organized in many sparsely urbanized (20 municipalities) and rural areas. These last, especially in bottom valleys, are devoted to intensive agriculture thanks to the presence of water and rich soils.

Because of the high elevation variability that characterizes the whole area, it is possible to distinguish different vegetation environments containing the most important forest formations typical of the southern Apennines Mountains (INEA, 2006). The prevalent types of forest are the deciduous oak wood and turkey oak wood.

In mountainous flat areas located above 1000 m, the thermophilic beech forest, typical of the southern Apennines, dominates especially in environments with appropriate levels of edaphic and atmospheric humidity. Noticeably, there is a considerable

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presence of chestnut coppice managed, mainly devoted to timber production and secondly to obtain fruits. Grass and pasture, as essential components of the Mediterranean agro-forestry systems, are widely scattered all over the region and ensure the maintenance of bovine livestock farming (<http://dati-censimentoagricoltura.istat.it/>, last access: 10 July 2014).

The unique and valuable environmental characteristics of this area come from the presence of the “Appennino Lucano, Val d’Agri, Lagonegrese” National Park and several protected areas (SIC – Site of Community Importance and SPA – Special Protection Area) with a wide range of flora and fauna biodiversity.

The wealthy fauna includes the occasional presence of wild boars and wolfs in addition to more common species adapted to anthropic landscapes and conditions such as fox, badger, porcupine and otter (Carone et al., 2012).

Both agricultural and forest sectors fulfil a quite important role within the economy of the valley. In the past, the economy was prevalently wood-based encompassing production of timber and non-wood products such as mushrooms and undergrowth fruits. More recently, we observed the escalation of interests and activities around the oil reservoir, which is the largest in Italy and considered as a strategic resource for the overall national energetic sector.

## 2.2 Data

We used satellite data acquired by the Thematic Mapper (TM) sensor flying on the Landsat 5 satellite.

In particular, we processed and analyzed two images available for the study area (188 path and 32 row GTCE), acquired on 10 August 1985 and 27 July 2009 with a spatial resolution of 30 m. A 24 year interval was considered suitable to capture significant changes in forest areas and to assess the evolution of composition and configuration of the investigated covers; dry-season imagery were selected for limiting the presence of cloud cover.

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Furthermore, a Forest map of Regione Basilicata was used as ancillary information (INEA, 2006; Costantini et al., 2011).

### 3 Methods

#### 3.1 Land cover maps and forest persistence

The two selected Landsat images were firstly preprocessed applying the transformation of digital numbers into radiance units and then into reflectance using the NASA-GSCF calibration coefficients (Chander and Markham, 2003; Barsi et al., 2007).

In order to detect and interpret the most important changes occurred in the investigated period, we produced land cover maps for the reference years. This was performed by using the multispectral clustering technique K-means (Richards and Jia, 1996; Brivio et al., 2006) that enables a preliminary partition of the area, then we refined the obtained classifications using training areas identified on the basis of spectral information and ancillary data derived from field surveys and forestry maps. Ultimately, we defined a complex ecomosaic of 14 classes for each year. From these, we derived also the following maps:

- Forest/Non Forest maps obtained by aggregating the fourteen classes in a binary map;
- Complex ecomosaic maps obtained by splitting forest classes into private and public ownership areas. This is the result of the intersection between the 14-classes ecomosaics and the available file of public forest boundaries.

Land cover changes occurred in the period examined were firstly pointed out by comparing the percentage of the landscape area occupied by each cover. In particular, as far the 14 classes ecomosaics, a transition matrix between initial and final distributions was evaluated to put into evidence alterations in the spatial patterns that synthesize the most important land cover trajectories emerging in the period investigated.

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To characterize forest areas we evaluated different fragmentation levels using a satellite-derived index of cover density ( $D$ ), based on NDVI (Normalized Difference Vegetation Index). NDVI is one of the most used vegetation indexes, considered as a proxy for plant biomass and vegetation activity (Rouse et al., 1974) and it is calculated as follows:

$$\text{NDVI} = \frac{\text{NIR} - \text{RED}}{\text{NIR} + \text{RED}} \quad (1)$$

where NIR and RED are respectively the near infrared and red channels of the electromagnetic spectrum corresponding to the bands 4 and 3 of the TM sensor.

In this study we derived NDVI from both the images and evaluated the relative density  $D$  for each cover class for both the years 1985 and 2009 according to the following formula:

$$D = \frac{\text{NDVI} - \text{NDVI}_{\min}}{\text{NDVI}_{\max} - \text{NDVI}_{\min}} \quad (2)$$

where  $\text{NDVI}_{\max}$  and  $\text{NDVI}_{\min}$  represent the maximum and minimum NDVI value for each forest class.

Afterwards, the values of the cover density index were classified into 4 ranges (< 20 %, 20–50 %, 50–80 %, > 80 %).

Integration, spatial analysis and any calculations were carried out in the GIS environment (QGIS 1.8, GRASS 6.4.2). The geographic information was projected to a common UTM zone 33 under the WGS84 datum.

### 3.2 Landscape metrics

The evaluation of the landscape structure was performed using metrics indicators, i.e., the core of the Landscape Ecology. Landscape metrics focus on the whole landscape considered as a mix of mutually interacting anthropic and natural covers and allow to analyze both landscape composition and configuration. Composition refers to the



variety and relative abundance of patch types within the ecomosaic and is typically quantified by diversity indices; configuration is linked to the spatial arrangement and the functional connectivity of the patches (McGarigal et al., 2009).

We computed landscape metrics from the land covers obtained according to the methodology described in the Sect. 3.1. This enabled us to determine the spatial relationships among patches within the ecomosaics and how the identified patterns change in the analyzed period (Jaeger, 2000). In particular, metrics greatly enhance our knowledge on how the growing impact of anthropogenic activities occurred in the study area during the reference period might have affected the ecosystem's ability to provide services and functions (Nagendra et al., 2004), facilitating also the interpretation of potential ecological consequences.

Moreover, in order to evaluate landscape properties under different forest management we applied landscape metrics to forest classes of the ecomosaics, differencing private and public ownership areas.

To this aim, we selected a minimally redundant and easily interpretable set of metrics indicators (Table 1), computed using public domain software FRAGSTAT 4.1 (McGarigal et al., 2012).

## 4 Results and discussions

### 4.1 Analysis of land cover transitions (1985–2009)

#### 4.1.1 Complex ecomosaic: 14 classes

Figure 2 shows the 14 classes-land cover maps derived from Landsat TM data for the years 1985 and 2009: 8 forest covers (Beech forests, Chestnut forests, Coniferous forests, Oak wood, Turkey oak wood, Manna ash and Hophornbeam, Riparian vegetation, Plantations including arboriculture plantations and reforestation with exotic species), 1 shrub-like cover (Shrublands), 1 herbaceous cover (Grass and Pasture

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areas), 3 non vegetated covers (Non Vegetated Areas including mainly artificial surfaces and bare soil, Burned Areas, Water Bodies), 1 Agricultural Area. Table 2 shows land cover classes and the corresponding acronyms adopted in this work.

At first glance it is evident that the underlying structures of the area have remained essentially unchanged during the investigated period (about 92 % of the total study area). This is confirmed by the transition matrix obtained by calculating the distribution of the identified fourteen classes for the investigated years (Table 3). The main diagonal, representing the proportion of each class that did not change through time, is populated by percentage values approaching to 100 % with the exception of Grass and Pasture (GP), Non vegetated Areas (NVA) and Plantations (PI).

In the analyzed period, the major change concerned Grass and Pasture (GP) areas. In fact, we observed the breakup of this cover (only 58 % remains unchanged) that significantly receded to the advantage of almost all the forest classes configuring a scenario of positive evolution of forest covers.

Covers benefiting most from this process are mainly Oak wood (Oa), Turkey oak wood (Toa), Coniferous forest classes (Co), Shrublands (Sh) and Riparian vegetation (Ri). The last one exhibits an enlargement that is small in absolute terms but very significant if compared with the limited size of the examined class in 1985 (about +18 %). Moreover also Beech forests (Be) show a slight increase and Chestnut (Ch) remain substantially stable.

Non Vegetated Areas (NVA) slightly expanded (about 100 ha). This is a low but worthy increase because it comes from the positive balance between two different phenomena: from one hand the growth of industrial areas (mostly coinciding with the building of the oil/gas pre-treatment plant in the municipality of Viggiano) to the detriment of Agricultural Areas (AA), on the other the conversion of some bare soil areas in prevalently natural covers.

The small loss of Agricultural Areas (AA, about 270 ha prevalently occupied by arable land) involving mountainous and steep sloping areas is well explained by the crisis of the agriculture sector in marginal areas and the consequent land abandonment as

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complex Mediterranean ecosystems (Simoniello et al., 2006). We also observe a reduction for NP that can be ascribed to the decrease of less densely vegetated areas within the main forest types between 1985 and 2009. On average, patches preserve the same degree of complexity with respect to the size (Shape) but extensiveness/compaction increases (Gyrate). Lastly, landscapes show for both the analyzed years a relatively appreciable level of connectedness (Cohesion).

At class level the results of the computed metrics are shown only for natural classes in Fig. 6.

The drop of NP and the notable growth of MPS express the reduction of the number of the patches and the increase of their sizes due to the forest transitions and the consequent rearrangement of the classes discussed in Sect. 4.1.1. The patch enlargement is not of similar magnitude in all the forest covers but, interestingly, it is twice or more only for Coniferous forests (Co), Shrublands (Sh), Manna ash and Hophornbeam (MH) and Riparian vegetation (Ri) classes.

Gyrate and Shape generally increase for all the forest covers except for Plantations (PI) that are characterized by very regular perimeters (Shape  $\sim 1.2$ ) coherently with the particular man-made nature of this cover that is prevalently composed of arboriculture plantations and reforestation with exotic species.

Furthermore, the decrease in NP consequently results in great changes that affect positively patch size and compactness (MPS and Gyrate indices), ultimately implying the increase in the distance between patches of the same class (ENN).

More broadly, the multitemporal analysis shows a tendency of forest areas to the expansion and the stabilization principally linked to the progressive disappearance of clearings within forest units (see in Fig. 7 two paradigmatic cases). This seems to be related to a more efficient management of cutting adopted in the last years and to the natural evolution of some forest types, particularly Coniferous (Co) afforested areas. Finally, in some areas we find an appropriate level of renaturalization due to the decrease of less densely vegetated areas with the expansion of mostly Beech forests (Be), Oak wood (Oa) and Turkey oak wood (TOa) covers.

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We computed the index of cover density ( $D$ ) to better highlight that fragmentation receded in the forest-dominated landscapes with the subsequent increase in vegetation density of forest covers (Fig. 8). The bar charts for the years 1985 and 2009 reveal a striking increase in vegetation cover density testified by higher values for the class  $> 80\%$  reached by all the forest covers, a substantial stability for the medium-high class (50–80 %, except for Coniferous forests (Co) and Shrubland (Sh) covers for which we observe an appreciable increase) and a considerable reduction for the first two vegetation cover density classes ( $< 20$  and 20–50 %).

### 4.2.2 Binary ecomosaic: Forest/Non Forest

On the basis of the simple landscape aggregation in Forest/Non Forest areas, the analysis performed at landscape level shows a substantial stability in diversity and evenness level (respectively SHDI and SHEI in Table 5). More precisely SHDI is characterized by quite low values as expected for very simple ecomosaics, whereas SHEI values can be considered very high, thus indicating a quite equal distribution in Forest and Non Forest classes.

We observe a marked decrease in the NP that is mainly caused by the natural evolution of afforested plantations (especially coniferous types) along with the choice of selective cutting in the broad-leaved forests and the natural evolution of herbaceous and shrub covers. Therefore, patches become larger and more compact (Gyrate) maintaining a very similar perimeter complexity (Shape). Lastly, the physical connectedness level (Cohesion) of the patches is constant over time and reaches very high values, close to the theoretical maximum of this metrics depicting a favorable scenario in terms of habitat quality for flora and fauna dispersal.

At class level histograms reported in Fig. 5 illustrate a picture of improved landscape ecological value principally because of the enlargement of forest areas covering about the 50 % of the whole study area in 2009 (PLAND) with respect to 1985 (about 45 %). This determined a consistent decrease in the NP, i.e. patches occupy a wider area ( $\sim 50$  ha, MPS) and take irregular forms (higher value for Shape index).

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These observed changes do slightly increase connectivity inside the patches of forest areas (Gyrate) that moves from about 70 m to about 80 m between the patch edges and the respective centroid, meaning a smaller exposure to external disturbance. Moreover also the ENN, computed between patches belonging to the same class, increases:  $\sim 5$  and  $\sim 10$  m respectively for forest and non forest areas.

The main insight that can be inferred from this analysis is the important reduction in 2009 of the fragmentation level affecting forest areas in 1985. Similar results have been obtained by evaluating the Shrubland class (Sh) separately from the forest areas, confirming also for this cover a significant decrease of fragmentation level.

### 4.2.3 Complex ecosaiic: private vs. public forest tenure regimes

Analyses were performed also for private and public managed forests by adopting the same metrics indicators of the previous elaborations at class level (see Figs. 9 and 10 showing the results of landscape metrics only for natural covers).

In general, metrics indicators put into evidence a different behaviour between private and public managed forest classes. The proportion of all the forest classes (PLAND) is increased in 2009 for both the analyzed forest tenure systems. However, in public domain larger increases are found for Coniferous forests (Co) that include several past and recent afforestation plantings. This proves the effectiveness of the implemented compensatory actions devoted to improve the ecological value of marginal or degraded areas. Differently, within private forest ownership Oak wood (Oa), Turkey oak wood (TOa), Shrublands (Sh) and Riparian vegetation (Ri) classes exhibit the highest rise in percentage composition. The growth of Riparian vegetation (Ri) class is mainly due to the application of protection regulations involving fluvial areas (issued in 1985) through the removal of agricultural land from the vicinity of riparian buffers. This enabled the establishment and/or evolution of vegetated buffer zones that show higher compaction levels (Gyrate) and a more complex patch perimeter (Shape). Remarkably, Riparian vegetation (Ri) shows an opposite trend within public ownership for the Shape index that decreases between 1985 and 2009, similarly to the behavior observed for Chestnut

forests (Ch) and Manna ash and Hophornbeam (MH). This last class, however, maintains an appropriate level of naturalness surpassed only by the corresponding value of beech forests (Be).

## 5 Conclusions

This study focuses on land cover changes and forest landscape evolution (1985–2009) in a typical Mediterranean agroforestry system (High Agri Valley) experiencing a rapid escalation of anthropic activities linked to the exploitation of the largest European on-shore oil reservoir.

We selected analysis tools suited for investigating the complex evolution of an area where the industrial acceleration contributed to strongly modify the socio-economic conditions with some repercussions on the landscape structure. These tools can be very precious to support policy makers to face the urgent need of harmonizing the unique natural heritage of the area with the demand of economic growth.

Land use trajectories and changes in landscape composition and configuration were captured by analyzing satellite-derived land cover maps. The main transformations concerned: the splitting of most of the grass and pasture areas to the advantage of all the forest covers; the enlargement of riparian vegetation that benefited from ad hoc regulations concerning habitat protection and conservation; the expansion of artificial areas due to the building of various infrastructures for oil and gas production and transport; the abandonment of less favoured agricultural areas located at high elevation.

We estimated a net forest increase of 11 % (about 5000 ha) during the time period 1985–2009 that coincides largely with the discovery and exploitation of the oil resources in the area.

Some of these quantitative findings about land cover variations confirm and integrate results obtained by independent studies concerning: the reduction in UAA (Utilised Agricultural Area) for all the municipalities involving particularly arable lands and permanent grass and pasture areas (Third and Sixth Italian Agricultural Census

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corresponding to the period 1982–2010, <http://dati-censimentoagricoltura.istat.it/>); the increase in forest areas reported by the Regione Basilicata (2009) that delineates a strong expansion of forests in the period 1960–2000 through an estimation based on the Corine Land Cover 2000 (Büttner et al., 2002) and the 1960 Land Use Map of Italy (Vitelli, 2007); the positive evolutionary tendency of the forest covers in the time span 1984–2010 highlighted by Mancino et al. (2014).

Forest covers do not appear to show macroscopic signs of anthropogenic disturbance. The use of landscape metrics indicators at both landscape and class level provided additional insight in understanding how forest areas have modified in the investigated period.

Briefly, all the forest classes increased their extent and have become generally more stable in functions and structure as a consequence of a larger patch size, a higher connectedness, a larger distance between similar patches and a more irregular perimeter. This analysis was carried out also taking into account the forest tenure regimes (public and private) to elucidate some particular aspects arising from the different type of management.

The most interesting cases are represented by Coniferous forests and Riparian vegetation classes. Part of the Coniferous forests was introduced since the Fifties for protecting steep slopes from slumps and landslides. Especially in public areas, these forests have reached healthy conditions and an encouraging level of naturalization suggesting that these interventions have been successful. The Riparian vegetation is characterized by a strong expansion and an appropriate arrangement very likely resulting from the application of legislations promoted in the past and recent years for protecting riparian vegetation (e.g., National Law 431/1985 on the protection of peculiar landscapes and environmental areas; EU Water Framework Directive 2000/60 – Integrated River Basin Management for Europe). This improvement of vegetation coverage along riverbanks activates a chain of positive effects including a better water quality, a higher ecological connectivity among river basins, and safer conditions for the biodiversity

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conservation of semi-aquatic species like otters (Morrow and Fischenich, 2000; Caranza et al., 2012; Carone et al., 2014).

More in general, the observed tendency to the expansion for all the forest classes, confirmed by the increase of the cover density index ( $D$ ), appears to be linked to the gradual reduction of small clearings within forests and to the internal forest dynamics entailing a recolonization of pastures and herbaceous areas.

The benefits arising from the expansion and consolidation of forest covers in the High Agri Valley could encompass biodiversity, ecological stability and biogeochemical cycles (IPCC, 2007).

Our results provide a coherent picture of the landscape evolution delineating not only the net increase/decrease balance of different covers but also how the area was transformed in terms of patch arrangement, size, connectivity and their corresponding ecological meanings.

Overall, considering the wide availability of satellite data characteristics (included the new EnMAP and Sentinel-2 mission), to respond to different requirements at the desired spatial and temporal resolution, the combined use of landscape metrics and satellite derived ecososaics represents a useful tool in operative contexts to optimize Forest Management Plans in different bioregional contexts. In particular, the obtained results corroborate the necessity to define specific operative landscape metrics guidance to assist policy and forest managers to fine-tune measures concerning silvicultural activities and refine them in itinere by performing multitemporal analyses. Future research could also investigate the real consequences induced by forest expansion in terms of effects on ecosystem services such as biodiversity, productivity, filtering air and water.

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**Table 1.** Landscape metrics indicators selected for evaluating the study area composition and configuration at landscape and class level.

Name	Description
NP	Number of Patches (NP): number of patches at the class or landscape level. $NP = \sum_{i=1}^n t_i$ where $n$ is the number of patch $t_i$ of a single class or of a landscape.
Gyrate	Gyrate: extent of each patch; i.e., how far across the landscape a patch extends its reach, given by the mean distance between cells in a patch. All other things equal, the larger the patch, the larger the radius of gyration. Similarly, holding area constant, the more extensive the patch (i.e., elongated and less compact), the greater the radius of gyration $GYRATE = \frac{\sum_{m=1}^k d_{ijm}}{k}$ where $d_{ijm}$ is the distance of each cell from the centroid of the patch $ij$ and $k$ is total number of cells that compose them.
Mean Patch Size	Mean Patch Size (MPS): the average area of patches comprising a landscape mosaic $MPS = \frac{\sum_{i=1}^n a_{ij}}{n_i}$ where $a_{ij}$ is the area (ha) of patch $ij$ ; $n$ is the patch number belonging to class $i$ .
Fract	Fractal dimension index (FRACT): describes the shape complexity across a range of spatial scales (patch sizes). It is bounded 1–2: it approaches 1 for shapes with very simple perimeters such as squares, 2 for shapes with highly convoluted, plane-filling perimeters. $FRACT = \frac{2 \ln(0.25 \rho_{ij})}{\ln a_{ij}}$ where $\rho_{ij}$ is the perimeter of patch $ij$ ; $a_{ij}$ is the area of the patch $ij$
Shape	SHAPE Index: is a measure of patch geometric complexity and it is given as the patch perimeter divided by the patch perimeter divided by the patch perimeter for a patch square of the same size. The index equals 1 for square patches of any size and increases without limit as the patch becomes increasingly non-square (i.e., more geometrically complex). $SHAPE = \frac{\rho_{ij}}{\min \rho_{ij}}$ where $\rho_{ij}$ is the perimeter of patch $ij$ ; $\min \rho_{ij}$ is the minimum perimeter of patch $ij$ in terms of number of cell edges; $n_i$ is the number of patch.
Cohesion	Cohesion Index is a measure of the physical connectedness of the corresponding patch type. It is an area-weighted mean perimeter-area ratio, invariant to changes in the cell size. It is bounded 0–100. Its value increases if the patch is more aggregated in its distribution, i.e. more physically connected; it is given as 0 if a class consists of a single patch. $COHESION = \left(1 - \frac{\sum_{i=1}^m \rho_{ij}}{\sum_{i=1}^m \rho_{ij} \sqrt{a_{ij}}}\right) \left(1 - \frac{1}{\sqrt{N}}\right)^{-1} \times (100)$ where $\rho_{ij}$ is the perimeter (m) and $a_{ij}$ is the area of patch $ij$ ; and $N$ is the total number of cells.
ShDI	Shannon's Diversity Index (ShDI) is a measure of diversity in community ecology, applied to landscapes. Its value is 0 when the landscape contains only 1 patch (i.e., no diversity); it increases as the number of different patch types increases and/or the proportional distribution of area among patch types becomes more equitable. $ShDI = -\sum_{i=1}^m (P_i \ln P_i)$ where $P_i$ is the proportion of the landscape occupied by patch type (class) $i$ .
ShEI	Shannon's Evenness Index (ShEI) is expressed as the ShDI divided by the logarithm of the number of patch types. ShEI = 0 when the landscape contains only 1 patch (i.e., no diversity) and approaches 0 as the distribution of area among the different patch types becomes increasingly uneven (i.e., dominated by 1 type); ShEI = 1 when distribution of area among patch types is perfectly even (i.e., proportional abundances are the same). $ShEI = \frac{-\sum_{i=1}^m (P_i \ln P_i)}{\ln m}$ where $P_i$ is the proportion of the landscape occupied by patch type (class) $i$ ; $m$ is the number of patch types (classes) present in the landscape.
ENN	Euclidean Nearest Neighbor (ENN) is the Euclidean measure of the distance (m) from a patch to the nearest neighboring patch of the same type, based on shortest edge-to-edge distance (from cell center to cell center) $ENN = h_{ij}$ where $h_{ij}$ = distance (m) from patch $ij$ to nearest neighboring patch of the same type (class)

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**Table 2.** Land cover classes and corresponding acronyms.

Land cover class	Acronym	Land cover class	Acronym
Beech forests	Be	Manna ash and Hophornbeam	MH
Oak wood	Oa	Riparian vegetation	Ri
Turkey oak wood	TOa	Plantations	PI
Chestnut forests	Ch	Grass and Pasture areas	GP
Coniferous forests	Co	Agricultural Areas	AA
Shrublands	Sh	Burned Areas	BA
Non Vegetated Areas	NVA	Water Bodies	WB

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**Table 3.** Transition matrix. Percentages values are computed on the area occupied by each class in 1985. Bold in the main diagonal represents the proportion of stable lands in the analyzed period (1985–2009) (abbreviations are shown in Table 2).

		2009														
Class		Be	Oa	Toa	Ch	Co	Sh	MH	Ri	PI	GP	AA	BA	NVA	WB	
1985	Be	<b>99.40</b>	0	0	0	0.30	0	0	0	0	0.24	0	0	0.06	0	
	Oa	0	<b>98.63</b>	0	0	0.13	0	0	0	0	1.01	0.23	0	0	0	
	Toa	0	0.01	<b>98.29</b>	0	0.09	0	0	0	0	1.44	0.10	0	0.07	0	
	Ch	0	0	0	<b>98.61</b>	0	0	0	0	0	1	0.24	0	0.15	0	
	Co	0	0.01	0	0	<b>99.29</b>	0	0	0	0	0.70	0	0	0	0	
	Sh	0	0	0	0	0.90	<b>97.13</b>	0	0	0	1.82	0	0	0.15	0	
	MH	0	0	0	0	0.32	0	<b>99.12</b>	0	0	0.56	0	0	0	0	
	Ri	0	0	0	0.01	0.06	0	0	<b>98.58</b>	0	1.01	0.34	0	0	0	
	PI	0	0	0	0	0	0	0	0	<b>95.94</b>	4.06	0	0	0	0	
	GP	0.88	4.32	8.92	1.06	7.81	12.49	1.82	3.04	0.65	<b>58.33</b>	0.44	0	0.23	0.01	
	AA	0	0.07	0.02	0	0	0.01	0	0.03	0.02	0.15	<b>98.25</b>	0	1.45	0	
	BA	5.83	2.96	1.71	0	8.89	18.58	0	0	0	57.99	4.04	<b>0</b>	0	0	
	NVA	0	0.19	0.49	0.09	0.23	0.11	0.05	10.92	0.01	4.30	3.26	0	<b>79.24</b>	1.11	
	WB	0	0	0	0	0	0	0	0	0	0	0	0	0	<b>100</b>	

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**Table 4.** Results of metrics applied at landscape level for the 14-classes ecomosaics for the years 1985–2009.

	SHDI	SHEI	NP	GYRATE_AM	SHAPE_AM	COHESION
LC_14cl – 1985	2.07	0.76	11 111	54.0	7.96	98.4
LC_14cl – 2009	2.11	0.80	6611	71.0	7.87	98.6



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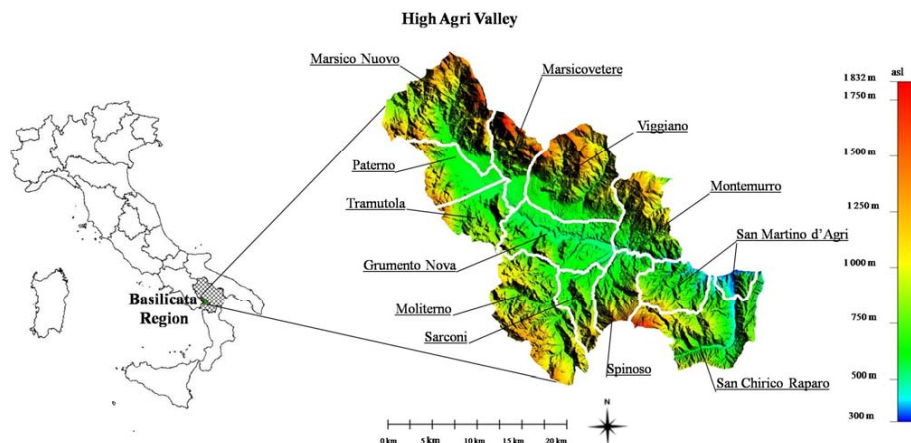
**Table 5.** Results of metrics applied at landscape level for the Forest/Non Forest ecosomaic for the years 1985–2009.

	SHDI	SHEI	NP	GYRATE_AM	SHAPE_AM	COHESION
LC_2cl – 1985	0.687	0.991	3250	65.152	23.808	99.721
LC_2cl – 2009	0.693	0.999	1960	83.654	24.247	99.772



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**Figure 1.** Location of the study area within the Basilicata region (southern Italy). The zoom shows the administrative limits of the municipalities of the High Agri Valley overlapped on the Digital Elevation Model (DEM) of the area provided by the Basin Authority of the Region at 20 m resolution.

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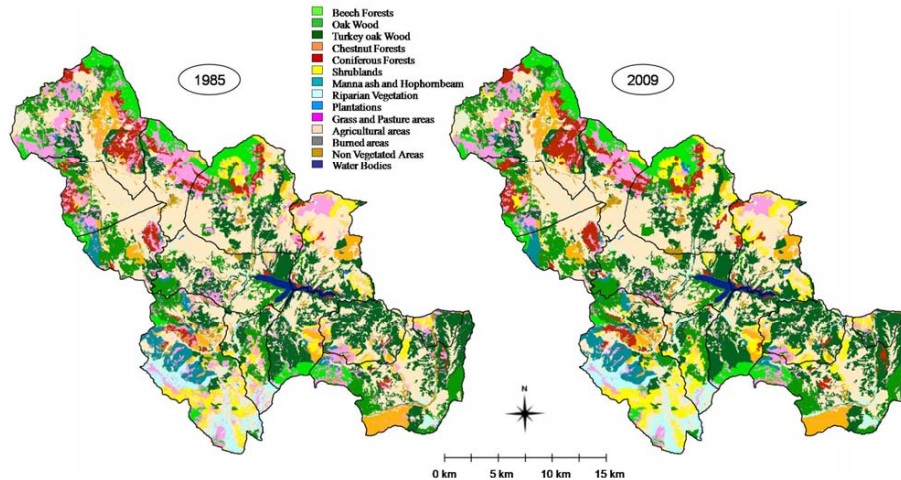
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**Figure 2.** Land Cover maps for the years 1985 and 2009 derived from Landsat TM data.

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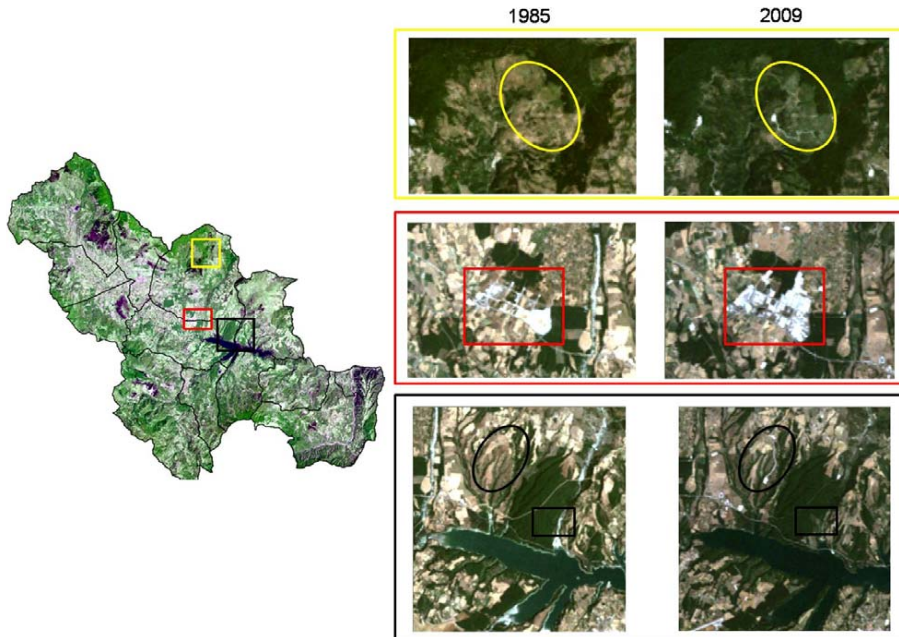
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**Figure 3.** Example of changes occurred in the High Agri Valley in the period 1985–2009. Areas highlighted on the map on the left are zoomed in on the right. In the yellow square an oil well and the service road with a wide herbaceous structure; in the red square the Viggiano industrial area where the Centro Olio Val d’Agri (COVA) is located; in the black square woodland variations close to Pertusillo Lake and new service roads.

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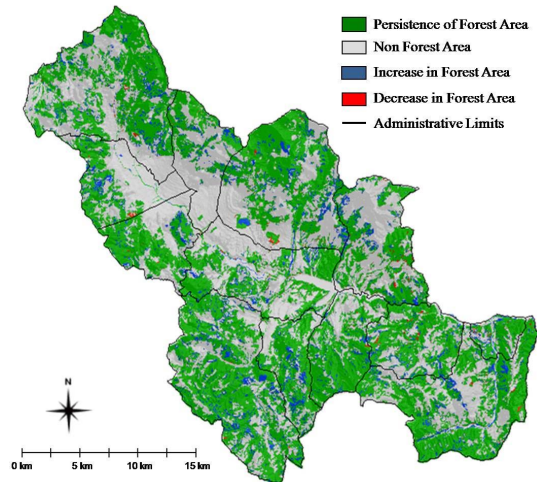
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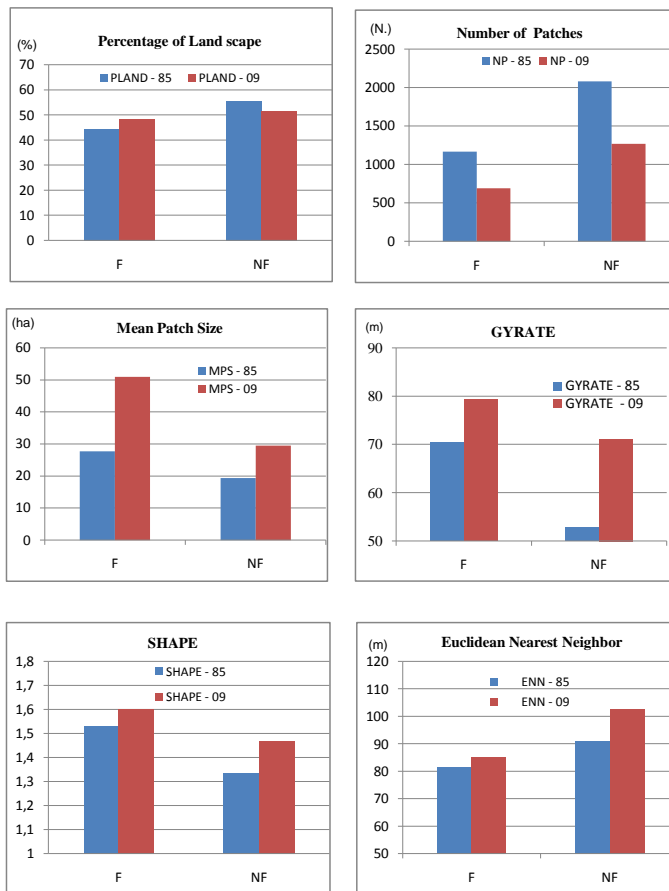


**Figure 4.** Forest area variation for the period 1985–2009.

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**Figure 5.** Comparison of the used metrics at class level for the Forest/Non Forest ecmosaic (1985–2009). For further details see Table 1.

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**Figure 6.** Comparison of the used metrics at class level for natural covers for the complex ecomosaics (1985–2009). For further details on landscape indicators see Table 1.

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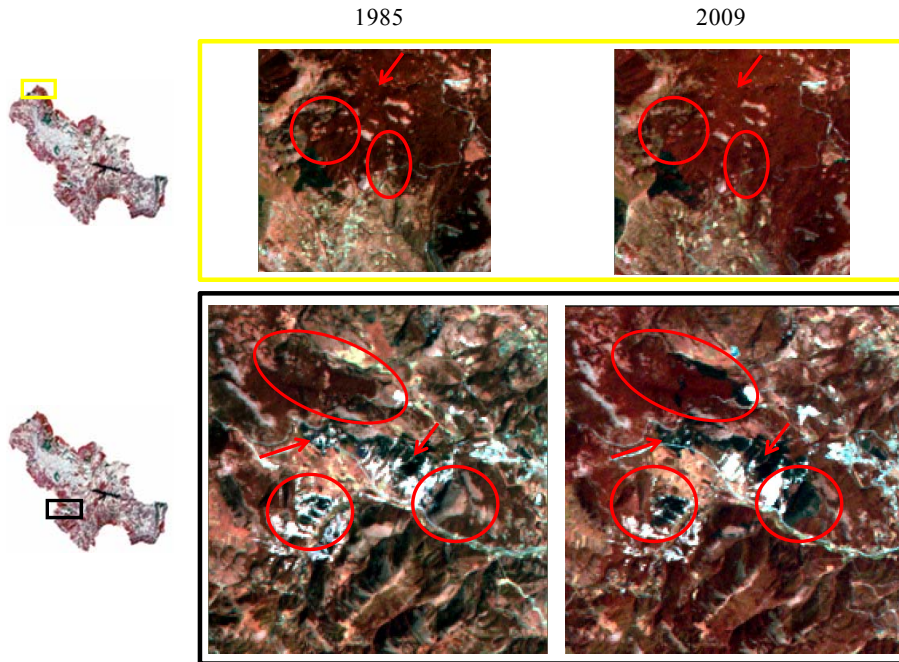
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**Figure 7.** Examples of changes detected from 1985 and 2009 in the High Agri Valley: reduction of pastures and sparsely vegetated areas due to natural forest evolution of Beeches (Be) on the top and Coniferous (Co) afforestation on the bottom. The RGB composition (R = band 5; G = band 2, B = band 1) of Landsat 5-TM enhances Broadleaved forests in dark red and Coniferous forests in dark green.

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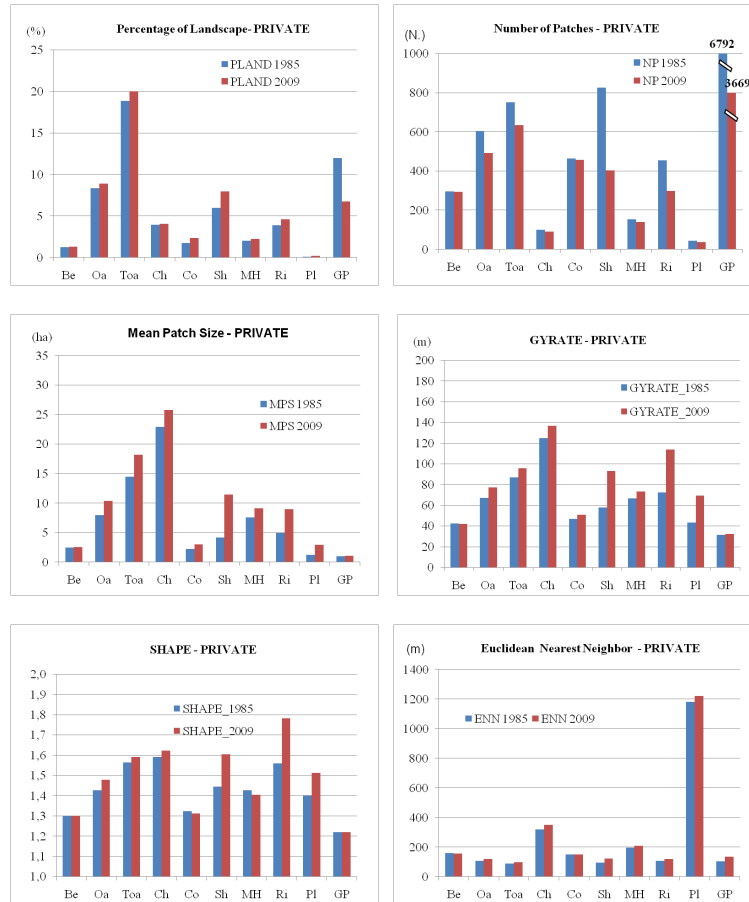
**Figure 9.** Comparison of the used metrics at class level for public ownership natural covers for the complex ecosasaics (1985–2009). For further details on landscape indicators see Table 1.

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**Figure 10.** Comparison of the used metrics at class level for private ownership natural covers for the complex ecosaiacs (1985–2009). For further details on landscape indicators see Table 1.

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